

Exxon Valdez Oil Spill
Restoration Project Final Report

Seasonal Movements of White-Winged Scoters (*Melanitta fusca*) from Prince William Sound,
Alaska

Restoration Project 273
Final Report
Volume II

This final report has been prepared for peer review as part of *Exxon Valdez* Oil Spill Trustee Council restoration program. Peer review comments from previous annual reports have been addressed in this final report.

Daniel H. Rosenberg
Michael J. Petrula
Douglas D. Hill

Alaska Department of Fish and Game
Division of Wildlife Conservation
525 W. 67th Ave.
Anchorage, Alaska 99518

December 2006

Seasonal Movements of White-Winged Scoters (*Melanitta fusca*) from Prince William Sound, Alaska

Restoration Project 273 Final Report Volume II

Study History: Restoration Project 273 continues studies initiated in 1998 (Rosenberg and Petrula 1999) and continued in 1999 (Rosenberg and Petrula 2000). The goal of this study is to monitor movements of white-winged scoters using satellite telemetry to increase our understanding of scoter ecology. White-winged scoters, a sea duck, are an important subsistence resource to the indigenous people of Prince William Sound. Scoter numbers, however, have reportedly declined for unknown reasons. Additional scoters were killed during the *Exxon Valdez* oil spill. The decline is a concern to both waterfowl managers and subsistence consumers. This report describes results of these studies.

Abstract: We used satellite telemetry to track the movements of white-winged scoters (*Melanitta fusca*) wintering in the Gulf of Alaska to determine timing and routes of migration and affiliations among wintering, breeding, and molting areas. This was the first successful satellite telemetry project to monitor the movements of white-winged scoters. We captured birds in Prince William Sound in April and May of 1999 and 2000. We monitored movements of 13 females and 9 males. Spring departure dates varied by year and breeding status. Five males and 11 females migrated to breeding areas in the 1) Yukon Flats National Wildlife Refuge, Alaska (n = 1 male); 2) Old Crow Flats, Yukon Territory (n = 2 females) and 3) Northwest Territories (n = 9 females, 4 males). We monitored 16 birds to coastal molting areas. Five birds went directly from wintering to molting areas. Molting areas were broadly distributed in the Beaufort Sea, Gulf of Alaska, and Bering Sea and primarily distinct from nesting and wintering areas. Eight of 12 breeding birds molted in the Beaufort Sea. Nine breeding birds (82%) returned to Prince William Sound. We identified an affiliation between wintering areas in the Gulf of Alaska and breeding and molting areas in the Northwest Territories.

Key Words: Alaska, Gulf of Alaska, Prince William Sound, Bering Sea, Northwest Territories, migration, breeding, molting, philopatry, population affiliation, satellite telemetry, scoter, sea duck, white-winged scoter.

Project Data: *Description of data* – Location and sensor data was recorded for each satellite transmitter. *Format* – Location and sensor data are in Microsoft Excel and DBASE IV spread sheet format. GIS coverage of Alaska and Canada showing scoter locations are presented in ArcView format. *Custodian* - Archived at ADF&G regional headquarters in Anchorage. Contact Dan Rosenberg at ADF&G, 525 W.67th Ave., Anchorage, Alaska 99518 (907-267-2453) (dan_rosenberg@fishgame.state.ak.us) or Mike Petrula (907-267-2159) (mike_petrula@fishgame.state.ak.us) for information. Project information can be viewed at http://www.wildlife.alaska.gov/index.cfm?adfg=waterfowl.scoter_home.

Citation: Rosenberg, D. H., M. J. Petrula, and D.D. Hill. 2006. Seasonal movements of white-winged scoters (*Melanitta fusca*) from Prince William Sound, Alaska. *Exxon Valdez* Oil Spill Restoration Project Final Report (Restoration Project /273), Alaska Department of Fish and Game, Division of Wildlife Conservation, Anchorage, Alaska.

TABLE OF CONTENTS

EXECUTIVE SUMMARY	vi
INTRODUCTION	1
STUDY AREA AND METHODS	3
Study Area	3
Equipment and Field Procedures	3
Data Acquisition and Analysis	5
RESULTS	6
Spring Migration.....	6
Departure from PWS.....	6
Migration Routes	6
Breeding Areas	8
Arrival Dates	8
Northwest Territories Core Breeding Area.....	8
Time at Breeding Areas	8
Molting Areas	9
Beaufort Sea.....	9
Bering Sea.....	10
Gulf of Alaska.....	11
Migration to Wintering Areas.....	11
Prince William Sound	11
Other Wintering Areas.....	12
Age Classes.....	13
Mortality and Transmitter Longevity	14
DISCUSSION	14
Spring Migration.....	15
Inland migration.....	15
Coastal migration	15
Breeding.....	16
Breeding Areas.....	16
Relationship to North American Waterfowl Population Surveys.....	16

Molting.....	17
Beaufort Sea.....	17
Bering Sea.....	18
Gulf of Alaska.....	18
Winter Fidelity.....	19
Mortality.....	19
CONCLUSIONS.....	21
Migration Patterns and Seasonal Use Areas.....	21
Migration.....	21
Breeding.....	21
Molting.....	21
Affiliations.....	22
Philopatry.....	22
Management and Research Implications.....	22
ACKNOWLEDGEMENTS.....	23
LITERATURE CITED.....	24

LIST OF TABLES

Table 1. General location of breeding, molting, and wintering areas.....	32
Table 2. Median departure date from PWS and arrival dates at breeding areas.....	33
Table 3. Median arrival and departure dates and duration at molting areas and arrival dates at wintering areas.....	34

LIST OF FIGURES

Figure 1. Map of Prince William Sound, Alaska showing capture locations in 1999 and 2000 for white-winged scoters implanted with satellite transmitters.	35
Figure 2. Movements of female white-winged scoters 20377, 30048, 20375, and 23894 marked with satellite transmitters from Prince William Sound, AK.	36
Figure 3. Movements of female white-winged scoters 23887, 23890, 27599, and 30049 marked with satellite transmitters from Prince William Sound, AK.	37
Figure 4. Movements of female white-winged scoters 23889, 23890, 27599, and 30049 marked with satellite transmitters from Prince William Sound, AK.	38
Figure 5. Movements of male white-winged scoters 12992, 29994, and 23888 marked with satellite transmitters from Prince William Sound, AK.	39
Figure 6. Movements of male white-winged scoters 30047 and 29992 marked with satellite transmitters from Prince William Sound, AK.	40
Figure 7. Movements of male white-winged scoters 30051, 20374, 24124 and 29995 marked with satellite transmitters from Prince William Sound, AK.	41
Figure 8. Map of breeding areas in Northwest and Yukon Territories, Canada used by white-winged scoters from Prince William Sound, Alaska.	42
Figure 9. Map of molting areas in the Beaufort Sea, Northwest Territories, Canada used by white-winged scoters from Prince William Sound, Alaska.	43

EXECUTIVE SUMMARY

We used satellite telemetry to track the movements of white-winged scoters (*Melanitta fusca*) wintering in Prince William Sound to determine timing and routes of migration and affiliations among wintering, breeding, and molting areas. Satellite transmitters allow researchers to determine large-scale movements of individual animals and link this information with life history data gathered at smaller spatial scales. This was the first successful satellite telemetry project to monitor the movements of white-winged scoters.

The migratory patterns of white-winged scoters are poorly understood. In North America, breeding white-winged scoters are widely dispersed over vast and remote areas throughout the boreal forests of Alaska and western Canada where they prefer to nest near permanent lakes. In western North America, post-breeding males and most females return to the coast to molt over a broad area from the Beaufort Sea to the Gulf of Alaska and west to the Aleutian Islands. The Pacific population winters in marine nearshore waters from eastern Kamchatka to Japan and from the Aleutian Islands, throughout the GOA, and south to Baja California.

Scoter populations in North America have been declining. Collectively breeding populations of all three North American species of scoters (white-winged, surf, black) have steadily declined by as much as 40% in Alaska and about 58% in North America since the late 1970's. Between 1973 and 1990, the estimated population of white-winged scoters wintering in Prince William Sound in the northern Gulf of Alaska declined over 80% from 24,000 to 3,300 birds. About 1,000 scoters (all species) died as a direct result of the *Exxon Valdez* oil spill in 1989. Population estimates for white-winged scoters in PWS are imprecise; apparently increasing since the spill but likely remaining well below historic levels. Scoters are an important food source for indigenous peoples of Prince William Sound and other areas affected by the *Exxon Valdez* oil spill.

Understanding the cause of these declines is hindered by a lack of information on the timing of movements and affiliations between wintering, breeding and molting areas. The underlying mechanisms for population change may occur anywhere in a bird's annual life cycle and we know little about seasonal movements and range affiliations throughout the annual cycle of white-winged scoters. This makes it difficult to effectively monitor population status and trends and develop conservation strategies.

We tested whether white-winged scoters from specific wintering areas are randomly distributed throughout the breeding and molting range or utilize more narrowly defined geographical "units" within the broader species distribution. Our objectives were to 1) identify breeding, molting, and staging areas, 2) identify affiliations among wintering, breeding, and molting areas, and 3) determine timing of movements and migration routes between seasonal use areas.

We captured 120 white-winged scoters in 2 areas of Prince William Sound in April and May of 1999 and 2000. Prince William Sound is characterized by fjord-like ports and bays surrounded by steeply rising mountains. Highly irregular in shape, it is approximately 160 km east to west and 140 km north to south.

We captured birds over open water using decoys and floating mist nets. In 1999, we implanted 13 birds with satellite transmitters (5 males, 8 females) and in 2000 we implanted 18 birds (9 males, 9 females). The transmitter was inserted into the coelomic cavity by an experienced veterinarian using standard aseptic surgical techniques including anesthesia. In 1999 surgery was performed on-site and birds were released within 2–5 hours of surgery. Responding to high rates of post-release mortality in 1999 we temporarily held birds pre- and post-surgery in outdoor pools at the Alaska SeaLife Center in Seward, AK in 2000. All birds were released at the capture sites.

We used Service Argos Inc., compatible PTT-100 implant transmitters. Transmitters were equipped with temperature and battery voltage sensors. The amount of data we received was limited by transmitter battery life. To conserve limited battery life transmitters were programmed for various on/off (duty) cycles based on predicted periods of seasonal movements. Signals were analyzed using Argos Data Collection and Location Systems processing. Locations were mapped using ArcView® GIS software. Individual transmitter performance (number of transmissions, location quality, and longevity) varied significantly, providing location data from 35 to 308 days (mean = 203, SD = 73).

We monitored movements of 13 females and 9 males. Departure dates varied by year and breeding status. Spring migrants departed Prince William Sound between 27 May and 3 June in 1999 (median = 31 May). In 2000 breeding birds departed between 18–27 May (median = 24 May) and nonbreeders between 14 June and 8 July. Birds used three widely separated breeding areas within the boreal forests of western Canada and Alaska. Five males and 11 females migrated to breeding areas in the 1) Yukon Flats National Wildlife Refuge, Alaska (n = 1 male); 2) Old Crow Flats, Yukon Territory (n = 2 females) and 3) Mackenzie, Anderson, and Horton River basins, Northwest Territories (n = 9 females, 4 males). Birds arrived on the breeding grounds between 25 May and 13 June. Birds arrived earlier in 2000 than in 1999. The core breeding area in the Northwest Territories encompassed approximately 45,000 km² in the open boreal forest. Females remained at breeding areas longer than males and longer stays may indicate that some possibly nested successfully. Nonbreeding birds took a coastal route, traveling directly to coastal molting areas in the Bering Sea or the northern Gulf of Alaska

We monitored 16 birds to coastal molting locations. Molting areas were primarily distinct and distant from nesting and wintering areas and distributed over a broad geographical area from the Beaufort Sea to the Gulf of Alaska and Bering Sea. Molting, like breeding, occurred in relatively few areas compared to the molting range of white-winged scoters in North America. Some females molted in breeding areas while others migrated almost 2,000km to discreet molting sites. The majority of breeding birds molted in the Beaufort Sea while a minority used the southern Bering Sea and northern Gulf of Alaska. Nonbreeding birds used only the southern Bering Sea and northern Gulf of Alaska.

We monitored the movements of 15 birds to wintering areas. Ten birds returned to Prince William Sound and one wintered nearby in Icy Bay in the Northern Gulf of Alaska. Icy Bay was the farthest east any bird wintered. At minimum three birds wintered on the Alaska Peninsula. A female wintered the farthest west, by Cape Sagak, Umnak Island, in the Aleutian Islands, about

1800km west of Prince William Sound. All but two birds migrated southwest from molting (or breeding) areas to wintering areas.

We documented strong evidence for winter site fidelity in breeding males and females. Of 11 breeding birds that we were able to monitor into the following winter, nine (82%) returned to PWS. All breeding males and 75% of breeding females returned. All birds ($n = 8$) that molted in the Northwest Territories returned to Prince William Sound. Nonbreeding birds had lower return rates and may represent dispersal. This is the first time winter site fidelity has been documented for white-winged scoters.

The majority of the core nesting area we identified in the Northwest Territories is surveyed annually as part of the Continental Waterfowl Breeding Population Habitat Survey conducted each May since 1955. Most scoter population estimates are derived from this survey. Prince William Sound birds used only three of these strata and primarily the eastern third of stratum 14 in the Northwest Territories. Numbers of scoters in stratum 14 have been declining since the late 1970's.

We identified affiliations of white-winged scoters from wintering areas with breeding and molting areas in Alaska and western Canada. Most significantly we identified an affiliation between Prince William Sound and a breeding area in the Mackenzie, Anderson, and Horton River basins in the Northwest Territories and a molting site in Liverpool Bay in the Beaufort Sea. This is the first time precise migratory connectivity and habitat affiliations have been documented for white-winged scoters. Breeding and molting areas were nearly distinct from those used by birds wintering farther south along the coasts of British Columbia and Washington State. More evidence is needed before we can solidify the spatial and temporal scale of this affiliation, but this type of association will help researchers and managers investigate, assess, or identify future resource related concerns. We also identified a link between molting birds near Cape Yakataga and Cape Suckling, the site of an historic scoter die-off, with staging and perhaps wintering areas in Prince William Sound.

This information will provide a basis for future population delineation and improved monitoring efforts, and allow managers to associate changing habitat and environmental conditions, harvest levels, and development proposals with distinct population segments. We present this as a small step toward identifying the mechanisms of population change. We also hope it will be useful to regional planners. Much additional effort among many disciplines is required before we can draw conclusions about population structuring at both smaller and larger spatial scales. However, we emphasize the importance of further defining and refining seasonal affiliations so we can comprehensively assess mortality factors throughout the annual cycle.

INTRODUCTION

We used satellite telemetry to track the movements of white-winged scoters (*Melanitta fusca*) wintering in the Gulf of Alaska (GOA) to determine timing and routes of migration and affiliations among wintering, breeding, and molting areas. Satellite transmitters allow researchers to determine large-scale movements of individual animals and link this information with life history data gathered at smaller spatial scales (Huntington et al 2004).

Three species of scoters breed and winter in North America: the Black Scoter (*Melanitta nigra*), Surf Scoter (*M. perspicillata*), and White-winged Scoter (*M. fusca*). Collectively, scoters are among the least studied of North American waterfowl (Godfrey 1986, Savard and Lamothe 1991, Henny et al. 1995) and relatively little is known about their life history, ecology, and seasonal distributions (Bellrose 1976, Herter et al. 1989, Goudie et al. 1994).

In North America, breeding white-winged scoters are widely dispersed over vast and remote areas throughout the boreal forests of Alaska and western Canada (Gabrielson and Lincoln 1959, Godfrey 1986) where they prefer to nest near permanent lakes (Brown and Brown 1981). Highest densities are in the Northwest Territories (NWT) of Canada and northeast Alaska (Palmer 1976, Bellrose 1980). As with other sea ducks, male scoters abandon incubating females in early summer and congregate at communal molting sites, which are often distinct from wintering areas (Salomonsen 1968). In western North America, post-breeding males and most females return to the coast to molt over a broad area from the Beaufort Sea to the GOA and west to the Aleutian Islands (Palmer 1976, Bellrose 1980, Johnson and Richardson 1982, Herter et al 1989, Alexander et al. 1988, Henny et al. 1995).

The migratory patterns of white-winged scoters are poorly understood (Alberta Sustainable Resource Development, 2002). Breeding birds from Saskatchewan migrate to both the Atlantic and Pacific coasts to winter (Houston and Brown 1983), but no clear demarcation exists on the breeding grounds between Atlantic and Pacific populations. Band recoveries suggest that birds breeding farther north and east are more likely to migrate to the Atlantic coast and conversely, those breeding farther south and west are more likely to migrate to the Pacific coast (Bellrose 1976). The Pacific population winters in marine nearshore waters from eastern Kamchatka to Japan (Kistchinski 1973) and from the Aleutian Islands, throughout the GOA, and south to Baja California (Palmer 1976).

While the general distribution of scoters has been documented, population estimates are hindered by the lack of differentiation among the three species during Continental surveys (Smith 1995). Collectively breeding populations of all three species have steadily declined by as much as 40% in Alaska and about 58% in North America since the late 1970's (Hodges et al. 1996, Canadian Wildlife Service 2000, S. Slattery, Ducks Unlimited Canada, pers. comm.).

Between 1973 and 1990, the estimated population of white-winged scoters wintering in Prince William Sound (PWS), in the northern GOA, declined over 80% from 24,000 to 3,300 birds (Klosiewski and Laing 1994). About 1,000 scoters (all species) died as a direct result of the *Exxon Valdez* oil spill in 1989 (John Piatt, pers. comm.). Population estimates for white-winged scoters in PWS are imprecise; apparently increasing since the spill but likely remaining below historic levels (Lance et al. 2001, Stephensen et al. 2001).

Understanding the cause of these declines and delineating the population segment(s) most affected is hindered by a lack of information on the timing of movements and affiliations between wintering, breeding and molting areas (Sea Duck Joint Venture Management Board 2001). The underlying mechanisms for population change (e.g. contaminants, habitat loss, harvest, food abundance, increased predation, climate change) are difficult to identify because they may occur anywhere in a bird's annual life cycle. In a migratory species, especially one with a broad range, identifying mechanisms of population change is confounded by a lack of specific information on affiliations between breeding, molting, and wintering areas (Henny et al. 1991, Webster et al. 2002), making it difficult to interpret surveys or the results of local research projects. Resource managers must define and focus efforts on specific geographic or biological "management units" (e.g. "subpopulations", "population segments") in order to identify constraints to population growth or viability (Moritz 1994). This requires knowledge of a species spatial and temporal distribution (migratory patterns) throughout annual cycles (Esler 2000).

Scoters are an important food source for indigenous peoples of North America, both currently and historically. Scoters comprise a significant portion of the waterfowl subsistence harvest in coastal and interior Alaska, including PWS (Birket-Smith 1953, Stratton 1981, Stratton and Chisum 1986, Scott et al. 1996, Wolfe et al. 1990.) and the Northwest Territories of Canada (Gwich'in Elders 1997). Bones from scoters are the most abundant avifaunal remains found at archeological sites in PWS over a 2,250-year period (Linda Yarborough, USFS, pers. comm.). In spite of their importance to the culture and subsistence lifestyles of the communities affected by the *Exxon Valdez* oil spill (Rosenberg and Petrula 1999, Huntington et al. 2004), we know little about seasonal movements and range affiliations throughout the annual cycle. This makes it difficult to understand and effectively regulate multi-season harvests across many jurisdictions.

Identifying migratory connectivity (Webster 2002) for white-winged scoters is necessary to interpret (or design) population surveys and a first step to detect causes of population change (Henny et al. 1991). A white-winged scoter die-off occurred during the molt near Cape Yakataga and Cape Suckling in the eastern GOA during 1990-1992 (Henny et al. 1995) but no definitive cause was identified. The lack of specific information on locations and movements of birds prior to their death increased the difficulty of identifying the cause of mortality (Henny et al. 1995) and the relationship to long-term population declines.

Scoters are difficult to band in large numbers and they have relatively low rates of harvest, making traditional banding studies impractical. There have been only 142 recoveries of 4,006 white-winged scoters banded in North America from 1923 to 2001

(Alisauskas et al. 2004). Most recoveries were at the banding location (Powell 2000) although bands from breeding areas in western Canada were recovered in the Atlantic, Pacific, and Mississippi flyways (Bellrose 1980, Houston and Brown 1983). Similarly, the wide range, remote locations, and at-sea habits of scoters make color marking and VHF telemetry impractical. Satellite telemetry offers the best approach to obtain information about seasonal distribution, migration routes, and timing of movements for birds that migrate over vast and remote regions.

We tested whether white-winged scoters from specific wintering areas are randomly distributed throughout the breeding and molting range or utilize more narrowly defined geographical “units” within the broader species distribution. Our objectives were to 1) identify breeding, molting, and staging areas, 2) identify affiliations among wintering, breeding, and molting areas, and 3) determine timing of movements and migration routes between seasonal use areas. In concert with other studies we hope this information will help develop a better understanding of population delineation and improve the design and interpretation of population surveys.

STUDY AREA AND METHODS

Study Area

Prince William Sound, Alaska (PWS) (ca. 60.5°N, 147.0°W), a large estuarine embayment of the northern GOA is characterized by fjord-like ports and bays surrounded by steeply rising mountains. (Fig. 1). Highly irregular in shape, it is approximately 160 km east to west and 140 km north to south. Tides can exceed 4.5m and water depth can reach 870m. Isleib and Kessel (1973) described the general physiography, climate, oceanography, and avian habitats of PWS. An estimated 3,000–24,000 white-winged scoters wintered in PWS since the spill (Stephensen et al. 2001).

We captured white-winged scoters in St. Matthew’s Bay (60.7°N, 146.4°W) from 26 April to 5 May 1999 and in Orca Inlet (60.5°N, 145.9°W) from 11–15 April 2000 (Fig. 1). Capture sites were about 37km northwest and 4km west of Cordova, Alaska respectively. Birds captured in St. Matthews Bay were attracted there by herring roe (spawn), which provided an abundant and easily accessible food source. Herring spawn was not present in Orca Inlet.

Equipment and Field Procedures

Birds were captured over open water using decoys and floating mist nets (Kaiser et al. 1995). Upon removal from nets birds were placed in small pet carriers with raised mesh liners and transported by skiff to a vessel anchored nearby. In 1999 we surgery was performed on the vessel. In 2000, birds were transported by float plane to the Alaska SeaLife Center (ASLC) in Seward, AK for surgery and observation. Responding to high rates of post-release mortality in 1999 (see Results); we held birds at the ASLC the following year. Birds were held for 5–9 days pre- and 12–21 days post-surgery.

In 1999 we implanted 13 birds (5 males, 8 females) and in 2000 we implanted 18 birds (9 males, 9 females). Surgical procedures were performed by an experienced veterinarian and followed protocols developed by Korschgen et al. (1996) with some modifications (Mulcahy and Esler 1999, Robert et al. 2000). At minimum, one veterinarian and one trained anesthetist performed the operation. Standard aseptic surgical techniques were practiced. Anesthesia was induced and maintained with isoflurane gas (Aerrane, Ohmeda) delivered in oxygen. Presurgical preparation included plucking feathers along the ventral midline (about 3.0 cm long by 2cm wide) between the distal end of the keel and the pubic bone to expose the incision site in the coelomic cavity (a practice that has since been discontinued). Another 1-cm² patch of feathers was plucked at a dorsal position nearest the intersection of the pubis and synsacrum and an incision was made for the antenna exit. The abdominal air sac was breached and the antenna was passed through a trocar inserted dorsally. The transmitter was inserted into the right abdominal air sac and all incisions closed with absorbable sutures. The transmitter was attached to the duck at a single point by a suture through the skin, body wall, and Dacron collar at the base of the antenna.

The large number of white-winged scoters we captured in PWS in 1999 and 2000 (n = 120) enabled us to generally select birds for satellite telemetry that were heavier, in definitive plumage (Palmer 1976) and, in most cases, appeared to be paired. We presumed that birds possessing these characteristics would be in superior condition and more likely to travel to breeding areas than lone individuals. Only birds aged as After Second Year (Gustafson et al. 1997) based on plumage characteristics (Palmer 1976) were selected for implants. In 2000, we measured bursal depth to further differentiate age-classes (Mather and Esler 1999). We divided birds into three age classes based on calendar year (Gustafson et al. 1997) following protocols of Mather and Esler (1999). If the bursa was absent or ≤ 3 mm we classified the bird as after-third-year (ATY). Birds with bursa depths from 4–10mm were classified as third-year (TY), and if bursa depth was >10 mm we classified the bird as second-year (SY) regardless of plumage.

Mean (\pm SD) body mass of female and male white-winged scoters implanted with transmitters was 1487.9 ± 107.0 g (n = 17) and 1671.8 ± 85.7 g (n = 14) respectively. Mean (\pm SD) body mass of all females and males captured was 1464.0 ± 118.1 g (n = 48) and 1648.3 ± 114.9 g (n = 71) respectively. Transmitter to body weight ratio ranged from 2.2 percent of body weight for the largest bird, an 1800g male, to 3.6 percent for a 1400g female.

Following surgery birds were returned to pet carriers and allowed to recover from anesthesia for 2–5 hours before release to the water within 0.5km of their capture site (1999) or returned to protected outdoor pools at the ASLC (2000). Once the latter birds remained dry at the ventral incision site; had normal hematocrits, total plasma solids, and leukocytes (buffy coat); gained mass, and exhibited no signs of trauma or lethargy, they were released at their capture sites.

We used Service Argos Inc., compatible PTT-100 implant transmitters (Microwave Telemetry Inc., Columbia MD). Transmitter design changed over the course of the study

due to changes in battery model and configuration. This resulted in transmitter weights varying from 50–51g in 1999 to 39–41g in 2000. Transmitter dimensions were 58mm long, 33mm wide, and from 9 to 14mm thick. A 21.6cm long Teflon-coated multi-strand stainless steel antenna exited from the posterior dorsal end of the transmitter and protruded 2cm before bending at a 90° angle. All satellite transmitters were reinforced to withstand external pressure and were equipped with temperature and battery voltage sensors.

Data Acquisition and Analysis

The amount of data we received was limited by transmitter battery life. Birds require small transmitters. This limits battery size, which limits transmitter life. To conserve limited battery life transmitters were programmed for various on/off (duty) cycles based on predicted periods of seasonal movements. All transmitters remained on for 6 consecutive hours. Off times varied from a minimum of 48 hours (spring migration) to a maximum of 120 hours (winter). We increased data collection during spring and summer at the expense of fall and winter.

Signals were analyzed using Argos Data Collection and Location Systems (Service Argos, Inc. Landover MD). We accepted all Argos Standard Location Processing with class codes 1, 2 or 3 (Service Argos 1996). When standard processing criteria were not met during a transmission cycle (generally fewer than 4 signals during a satellite overpass), we used Argos Auxiliary Location Processing (class codes A, B). We removed aberrant locations associated with Class codes A, B, and 0 which is typical for data obtained through the Argos ‘Doppler’ system (Fancy et al. 1988) by using qualitative and quantitative criteria based on travel distance, travel rate, and redundancy from previous or subsequent locations (Ely et al. 1997, David Douglas USGS filtering algorithm). Through this filtering process we attempted to eliminate all locations lacking biological plausibility (Ely et al. 1997). Locations were mapped using ArcView® GIS software. Birds that died (verified by temperature sensor) or stopped transmitting signals within 30 days of release were not included in the analysis.

We presumed seasonal status (spring departure, breeding, molting, wintering, migration) by correlating dates and patterns of movement with geographical locations and published information, historical accounts, and surveys. A breeding area was assigned to a bird if it spent a minimum of 9 days at an interior location during the breeding season before departing to a molting area. If a bird spent time at two or more locations we used the latter. It was not intended as an indication of breeding status.

We used the best location in each duty cycle for mapping and distance calculations. We selected from all plausible locations to determine migration routes. We estimated dates of spring migration from PWS by using the median date between two sequential locations (last date in PWS and first date outside PWS) and rounding towards the first location in the sequence. We followed this pattern throughout migration (e.g. median date between last location at breeding area and first location after breeding area to indicate last date at breeding area). We did not estimate arrival and departure dates between seasonal use

areas if we did not receive a location for a minimum of 8 consecutive days between movements.

Methods used in this study were approved by the *Exxon Valdez* Oil Spill Trustee Council and the Institutional Animal Care and Use Committee of the Alaska SeaLife Center. Capture and handling of birds followed Ornithological Council guidelines (1997).

RESULTS

We captured 120 white-winged scoters. We monitored movements of 13 females and 9 males from PWS (Table 1). Individual transmitter performance (number of transmissions, location quality, and longevity) varied significantly. Individual transmitters provided location data from 35 to 308 days (mean = 203, SD = 73) and the total number of locations per individual ranged from 30 – 814 (mean = 349.9, SD = 250.4). In 1999, we received 2,339 locations from 7 birds and in 2000 we received 5,358 locations from 15 birds. We used 16.3% of all locations in our analysis. Sample size at a given staging, breeding, molting, or wintering area varied throughout the study as transmitter performance and individual migration patterns varied. We were not able to track a bird throughout a full year, but we were able to track most birds to breeding and molting areas and back to wintering areas (Table 1).

Spring Migration

Departure from PWS

All 22 birds departed PWS. We observed two different migration patterns in spring. Breeding birds traveled inland, departing PWS in a northeasterly direction with the majority of birds migrating to breeding areas in the NWT, Canada (Fig. X). Nonbreeding birds from PWS took a coastal route, traveling westerly or easterly to coastal molting areas in the Bering Sea of western Alaska or the northern GOA (Fig. X).

Departure dates from PWS varied by year and breeding status (Table 2). Spring migrants departed PWS between 27 May and 3 June in 1999 (median = 31 May). Departure dates for the two males not migrating to breeding areas were within the range of the five females that migrated to breeding areas. In 2000, 12 of 15 birds departed PWS between 18 – 27 May (median = 24 May) and the remaining three birds (nonbreeders) between 14 June and 8 July (Table 2). Unlike earlier migrants that went to breeding areas, the latter three birds migrated to coastal molting areas (see *Coastal Routes*, below).

Migration Routes

Spring migration to breeding areas was fairly rapid and direct. Transmitters were programmed to receive location data six out of every 54 hours, affording us few opportunities to monitor birds during migration and identify routes and brief stopover points. In addition we received fewer messages and poorer quality messages during periods when birds were flying.

We received enough locations from seven birds to suggest a general migration route. Birds migrated northwest from PWS, through the Copper River basin, circumnavigating the northwest side of the Wrangell Mountains via the upper Copper River Valley and proceeding east and crossing the Alaska Range through the Nabesna River Valley or Mentasta Mountains to wetlands in the upper Tanana River Valley (Tetlin). A female died during spring migration near the outlet of Tetlin Lake soon after her estimated arrival on 28 May 1999.

Old Crow and Yukon Flats. Birds heading to breeding areas in the Yukon Flats and Old Crow proceeded north from Tetlin (Table 1). Three birds arrived east of Ft. Yukon (66.6°N, 145.3°W) by 29 May (Figs 2A, 3D, 5C), two days later than observed in 1962, a year when the Yukon River breakup was later than normal (Lensink 1962). Birds staged for about one week in the vicinity of the confluence of the Black and Porcupine Rivers in the Yukon Flats. We cannot determine the route from Tetlin to the Black River, though birds likely crossed the Yukon-Tanana uplands (Wahrhaftig 1965), a minimum flight distance of about 410km or proceeded northeast into the Yukon River drainage and then followed it to the northwest. From here birds proceeded further east on the Yukon Flats (Fig. 5C), or followed the Porcupine River to the Old Crow Flats (Fig. 2A, 3D) (Irving 1960)

Northwest Territories. Birds migrating to the NWT (Table 1, Fig. 2B–D; 3A,B,D; 4A–D; 5A–C; 6A,B) traveled east from Tetlin to the Stewart River Valley in the Yukon Territory, then northeast between the Ogilvie and Wernecke Mountains to the Peel River basin. From here birds proceeded east, crossing the Mackenzie River near Fort Good Hope (66.3°N, 128.6°W). We have several scattered locations between the Peel and Mackenzie Rivers in late May prior to birds moving further east to breeding areas.

Coastal Routes. Four males and one female went from wintering to molting areas without stopovers at breeding areas (Figs. 3C, 7A–D). The migration of nonbreeders was generally more prolonged than the rapid and direct migration of birds to breeding areas. All nonbreeding birds migrated to the southern Bering Sea or northern GOA. Individual migrations varied. A male staged in lower Cook Inlet (Kamishak Bay) before traveling further west to Togiak Bay in the southern Bering Sea (Fig. 7C). A second male remained in lower Cook Inlet to molt (see Molting Areas below) (Fig 7B). The female traveled west through Cook Inlet but sporadic data prevented us from being more detailed about the timing and locations of her movements. By 31 July she had arrived along the west coast of the Nushagak Peninsula (Fig. 3C). The fourth bird, a nonbreeding male, took the most direct route to the molting area, migrating from PWS to Cape Newenham within six days (Fig 7A). Only one bird migrated east. This nonbreeding male took the shortest migration route of any bird (226km), flying from PWS to a staging area at the mouth of Icy Bay in the northeastern GOA (Fig. 7D).

Breeding Areas

Five males and 11 females from PWS migrated to breeding areas in the 1) Yukon Flats National Wildlife Refuge, Alaska (Yukon Flats) (n = 1 male); 2) Old Crow Flats, Yukon Territory (n = 2 females) and 3) Mackenzie, Anderson, and Horton River basins, NWT (n = 9 females, 4 males) (Table 1, Fig. 8). Although we have no empirical evidence of our transmitted birds actually nesting, our satellite locations are within known breeding areas (Palmer 1976, Bellrose 1980, Brown and Fredrickson 1997).

Arrival Dates

Birds arrived on the breeding grounds between 25 May and 13 June (Table 2). Birds arrived earlier in 2000 than in 1999. The estimated median arrival date at breeding areas in 1999 for four females was 6 June. The estimated median arrival date in 2000 was 31 May for seven females and five males (Table 2). The two Old Crow breeders arrived between 5–8 June, later than Irving (1960) observed in 1957 and likely due to a latter spring thaw in 2000 (S. Haszard, pers. comm.). The only Yukon Flats breeding bird, a male, arrived on approximately 2 June, preceding the dates for mean nest initiation (Safine 2005). We do not have historical data from the NWT breeding areas to compare with our estimated dates of arrival.

Northwest Territories Core Breeding Area

PWS birds used breeding areas primarily in the lower Mackenzie and Great Bear sub-basins of the Mackenzie River (Mackenzie River Basin Board 2004) and further north in the basins of the Horton and Anderson Rivers (Table 1, Figs. 2B–D; 3A,B,D; 4A–D; 5A–C; 6A,B). The core breeding area in the NWT encompassed a triangular area of approximately 45,000 km² in the open boreal forest, with the triangle tapering in width from south to north (Fig. 8). The area was bordered to the southeast (65.7°N, 123.3°W) by the Great Bear Lake, to the southwest (65.9°N, 127.6°W) near Fort Good Hope and the Mackenzie River and to the north (68.9°N, 127.5°W), west of Sadere Lake between the headwaters of the Anderson and Horton Rivers. The size of this core breeding area is approximately 3 times the area of PWS. All breeding birds in the NWT were within the boundaries we described for the core breeding area with one exception, a female, who “nested” 175km to the west (Fig. 2D).

Time at Breeding Areas

As is typical for sea ducks, females remained at breeding areas longer than males (Table 2) and longer stays may indicate that some possibly nested successfully. White-wing scoter females commonly combine broods (crèching) when ducklings are less than 1 week old (Brown and Brown 1981) and the combined egg laying and incubation period is from 37–44 days (Brown 1977, Brown 1981). Thus, five females were present at a breeding site long enough (minimum 62 days) to have nested successfully (Figs. 2B,C; 3A,D; 4D). A sixth female was at the breeding site for 47 days (Fig. 2A). Four of these females used the core breeding area in the NWT and two used Old Crow Flats.

We permanently lost contact with 2 females while at breeding areas after 39 and 62 days. We do not know if the birds died or the transmitter failed. Body temperatures were normal in both birds when we received our last transmission of data.

Molting Areas

We were able to monitor 16 birds to coastal molting locations (Table 3). Molting areas were primarily distinct and distant from nesting and wintering areas and distributed over a broad geographical area. We divided molt chronology between birds that migrated to molting areas from breeding areas and nonbreeders that migrated from wintering to molting areas without stopovers at breeding areas (Table 3). We identified three general areas used by molting scoters. The majority of breeding birds molted in the Beaufort Sea while others used the southern Bering Sea and northern GOA (Table 1, Fig 9). Nonbreeding birds used only the southern Bering Sea and northern GOA (Figs. 3A–D, 7A). Nonbreeding birds generally arrived at molting sites earlier than breeding birds (Table 3). In addition to coastal molting sites, two females molted on breeding lakes in the NWT (Figs. 2B, 4D).

The time each individual spent at breeding areas varied greatly, thus the advent of molt migration varied for each individual and occurred over an extended time period (Tables 2 and 3). Once initiated, molt migration was fairly direct and most birds generally arrived at molting sites within 9 days of departing breeding areas with some exceptions (see below). Birds with distinct molting areas (i.e. separate from breeding or wintering areas) spent an average of 83 days at molting areas (Table 3). Following molt, scoters generally remained at or near the same location before migrating to wintering areas.

Breeding birds from the NWT tended to molt in the NWT (Beaufort Sea) but we did not observe any absolute affiliations between breeding and molting areas. In addition to the Beaufort Sea, NWT birds molted in the northern GOA and the southern Bering Sea (Nelson Lagoon, 56.0°N, 160.8°W). The two females from Old Crow dispersed southwesterly to molting sites in the northern GOA (PWS) and the southern Bering Sea (Kuskokwim Bay). The lone Yukon Flats bird, a male, migrated northeasterly to join NWT birds in the Beaufort Sea.

Beaufort Sea.

Eight of twelve breeding birds molted in the Beaufort Sea, including seven of nine birds that nested in the NWT (Table 1, Fig. 9). The primary molting site for breeding birds was Liverpool Bay ((70.3°N, 128.4°W) off the coast of the NWT (3 Females, 2 Males) (Fig 9). Another male and female were present in Liverpool Bay but we lost data reception before we could confirm molting. An additional male molted to the east in Franklin Bay near the northern tip of the Parry Peninsula (70.1°N, 124.5°W) (Figs. 5B, 9). This was the most easterly molting site used by one of our birds. All birds that molted in the Beaufort Sea arrived from breeding areas in the NWT with one exception. The lone male from the

Yukon Flats also molted in Liverpool Bay (Fig. 5C). Nonbreeding birds did not molt in the Beaufort Sea.

Molt migration from breeding areas was relatively short and direct though a few birds staged at tundra and coastal locations before arriving at molting areas. A female migrated to the Beaufort Sea coast at Darnley Bay (69.6°N, 123.9°W) before moving 170km west to Liverpool Bay (Fig. 4C). Darnley Bay was the most easterly coastal staging area used by one of our white-winged scoters.

Two birds, 1 female and 1 male each spent about two weeks in the core breeding area near Lac Maunoir (67.5°N, 125.0°W) in the headwaters of the Anderson River from late May to early June before moving approximately 320 km northwest to coastal tundra locations on the Nicholson and Tuktoyaktuk peninsulas respectively (Figs. 4A, 5A). The female spent 32 days southeast of Nicholson Peninsula between Wood Bay and Liverpool Bay before moving to a molting area in Liverpool Bay. The male spent three weeks on the tundra before we lost all contact (lost transmission). During that period we documented one brief round trip to a known molting site in Liverpool Bay. Arner et al. (1985) observed an influx of white-winged scoters on the Tuktoyaktuk Peninsula in late June but observed no evidence of nesting. A lack of nesting on the Arctic coastal plain of western Canada was supported by Porslid (1943) and Salter et al. (1980). Thus, we don't believe birds attempted to nest at these locations but staged before continuing their migration to molting areas, most likely in Liverpool Bay.

Satellite data confirmed the presence of a female in Liverpool Bay on 2 October and based on her next location she possibly remained until 5 October. This is the latest departure date we can reliably estimate for the Beaufort Sea.

Bering Sea

Five birds molted in the southern Bering Sea. Birds arrived in the Bering Sea from wintering and breeding areas. Molting sites were widely dispersed from Kuskokwim Bay on the north to Nelson Lagoon in the southwest a direct distance of about 450km.

Two breeding females from the Anderson River drainage, NWT, and Old Crow Flats molted in Nelson Lagoon and Kuskokwim Bay respectively (Figs. 3A,D). The molt migration from the Anderson River drainage to Nelson Lagoon was the longest we identified. This bird flew the 1903km from the breeding area to the south side of the Alaska Peninsula within 72 hours. After staging for at most one week, just northeast of Yantarni Bay, it flew another 303km northwest across the Alaska Peninsula to Nelson Lagoon. Nelson Lagoon is the most southwesterly molting site we identified and is approximately 2,342 km from our most easterly site in Franklin Bay. The Old Crow female remained 111 days in Kuskokwim Bay, the longest a breeding bird spent at a coastal molting site that was distinct from its wintering area. She finally departed the molting area after 27 November for a wintering area further west in the Aleutian Islands (see Winter Areas, below).

Three birds, all nonbreeders, molted in disparate locations from Cape Newenham (north side) to the west coast of the Nushagak Peninsula (Figs. 3C, 7A,C). After molting at Cape Newenham the bird moved north into Kuskokwim Bay where it remained through 29 November when we lost contact. The bird spent a minimum of 156 days in the Bering Sea from its arrival at Cape Newenham about 27 June until we lost contact. Unfortunately we do not know where it wintered.

Gulf of Alaska

Two breeding birds and two nonbreeding birds molted in the northern GOA. Molting areas were widely dispersed from Cook Inlet to Cape Yakataga (Figs. 2A, 6A, 7B,D). A female from the Old Crow Flats molted in Knowles Bay in northeastern PWS (Fig X). It arrived in Knowles Bay on approximately 26 July and remained until 24 September. A male from the NWT migrated to Cape Suckling by 10 July. We next located it in Controller Bay, 60km northwest, in late August (Fig. 6A). We lost contact (i.e. did not receive data) with this latter bird for over 6 weeks and could not confirm where it molted. We believe it likely molted in the northern GOA, probably near Cape Suckling, a known molting area, (Henny et al. 1995), based on July and August locations. Two nonbreeding birds, both males, molted at Cape Yakataga (Fig. 7D) and lower Cook Inlet (Chinitna Bay) (Fig. 7B) respectively. The former bird remained at Cape Yakataga from 20 June through 28 August and the latter bird resided in Chinitna Bay from 13 June until 27 August.

Migration to Wintering Areas

We monitored the movements of 15 birds to wintering areas (Tables 1 and 3). Birds not included in Table 3 had the same molting and wintering areas. Ten birds returned to PWS and one wintered in the Northern GOA (Icy Bay) (Table 1) (Fig. 7D). Icy Bay was the farthest east any bird wintered. At minimum three birds wintered on the Alaska Peninsula (see Other Wintering Areas, below), two on the north side (Bering Sea) and one on the south side (GOA). A female wintered the farthest west, by Cape Sagak, Umnak Island (52.8°N, 169.2°W), in the Aleutian Islands (Fig. 3D). Umnak Island is about 1800km west of PWS. All but two birds migrated southwest from molting (or breeding) areas to wintering areas.

Prince William Sound

Breeding birds (birds that migrated from PWS to breeding areas) exhibited a high rate of return to PWS. Of 11 breeding birds that we were able to monitor into the following winter, nine (82%) returned to PWS (Table 1). All breeding males returned while 75% of breeding females returned. All birds (n = 8) that molted in the NWT returned to PWS (Table 1).

The first breeding bird returned to PWS on 25 July (estimated – see methods). This female, returning from Old Crow Flats was the only bird to molt in PWS (Knowles Bay) (Fig. 2A). Birds molting in the NWT returned much later. The first bird to return from the

NWT, a female, arrived on 15 September after molting on a breeding area lake. The first coastal molting bird, a male from the Beaufort Sea, returned on 21 September. Excluding the first bird that returned in July, the average date of return to PWS was 27 September and the median date was 1 October. All birds returned by 11 October. We estimated peak return occurred between 29 September and 7 October.

The return migration to PWS was generally rapid and direct, usually occurring within a few days to a week. Birds from the NWT all returned via inland routes, following river valleys within the Yukon and Mackenzie River basins. Our data suggests three birds possibly returned to PWS from the NWT by traveling on the eastside of the Wrangell Mountains but location data were not sufficient. Return routes likely varied among flocks.

In 1999, a female possibly remained in Liverpool Bay until 5 October (see Molting Areas, above). It was next located in the Black River drainage in the interior of northeast Alaska (66.1°N 142.6°W) the evening of 5 October before returning to PWS by 8 October. This bird migrated 1,334km from the Beaufort Sea to PWS in a maximum of 147 hours and possibly within 48 hours (Fig. 3B). Another female, in what we believe is atypical, returned to PWS via southeast Alaska after having molted on a freshwater lake in the NWT (Fig. 2B). She departed the NWT sometime after 14 September and traveled through the upper Mackenzie River Valley to Lynn Canal, arriving no later than 19 September. She staged for about two weeks at the south end of Baranof Island before returning to PWS within the first few days of October.

Among five nonbreeders, only one bird, a male, returned to PWS (via Cook Inlet) on 4 September after molting in Togiak Bay (Fig. 7C). This bird returned before all but one breeding bird (see above). This was the only bird to migrate east from a molting to wintering area. Thus it was the only bird to return to PWS among the two breeders and four nonbreeders molting from Cook Inlet west (Table 1).

After returning to PWS birds dispersed from Sheep Point east to Glacier and Perry islands and south to Green Island (Figs. 2A,B; 4A,D; 5B, 7C).

Other Wintering Areas

Icy Bay in the northern GOA was the most easterly wintering location for one of our birds (Fig. 7D). Prior to moving to Icy Bay this nonbreeding male molted at Cape Yakataga then moved west and staged from mid-September to early December outside of barrier islands in the Copper River Delta. It arrived in Icy Bay in mid-December and remained there until mid-February 2001 when we lost contact.

Three birds wintered on the Alaska Peninsula (Figs. 3A,C; 7B). We lost contact with another bird, a male, while still in Kuskokwim Bay on 29 November 2000 (Fig. 7A.). We suspect, if still alive, it wintered in the Bering Sea or further west. A female, having molted in Port Moller remained there until we lost contact in late November 1999 (Fig.

3A). We tentatively identified this as her wintering area but she may have moved subsequently.

The second female, a nonbreeder wintering on the north side of the Alaska Peninsula, moved among several locations throughout the winter (Fig. 3C). She was one of two birds without a definitive southwesterly migration from the molting to wintering area. After molting by the Nushagak Peninsula she moved to the north side of Hagemester Island (mid-October to mid-December, 2000), then returned to her molt site, remaining until early January 2001. She then proceeded west along the Alaska Peninsula by Port Moller where she remained until early February before moving northeast to the mouth of Ugashik Bay where she remained until early March 2001. Our last location before losing contact was of poor quality, but plausible and intriguing. It placed the bird on Becharof Lake, on the Alaska Peninsula, southeast of Ugashik in mid-March. We will never know if this bird was migrating east, perhaps back to PWS or continuing a series of winter movements. She illustrates the fascinations and frustrations inherent in satellite telemetry studies.

One bird wintered on the south side of the Alaska Peninsula in Chiginagak Bay (Cape Providence) (Fig. 7B). After molting and staging in lower Cook Inlet, this nonbreeding male traveled west along the south side of the Alaska Peninsula spending two months in the fall at Puale Bay before arriving in Chiginagak Bay in late October. We lost contact with this bird in mid-December while still in Chiginagak Bay. We identified this as its wintering area, recognizing the possibility of additional winter movements.

Finally, a female that bred near Old Crow and molted in Kuskokwim Bay, wintered the farthest west, by Cape Sagak, Umnak Island (52.8°N, 169.2°W), in the Aleutian Islands (Fig. 3D). She departed Kuskokwim Bay after 27 November, migrated west by Port Moller on 30 November and arrived at Umnak Island the first week of December. She remained there until 28 January 2001 when we lost contact.

Age Classes

Distribution of age classes based on bursal depth among the eight females and seven males we monitored from PWS in 2000 were as follows: After Third Year (ATY) n = 6 Females, 5 Males; Third Year (TY) n = 1 Female, 1 Male; and Second Year (SY) n = 1 Female, 1 Male.

All but one bird classified as ATY went to a breeding area. A female, the one exception, flew directly to a molting area in western Alaska. Of the two TY birds, the female migrated to a breeding area while the male flew directly to a molting area (Figs. 4A, 7A). Of the two birds classified as SY, the female migrated to a breeding area in the NWT where she remained to molt, staying until mid-September (Fig. 2B). The male flew directly to a molting area east of PWS (Fig. 7D). Both birds classified as SY by bursal depth had acquired definitive plumage (Palmer 1976). The bursa depths of two TY birds were at the lower range for that category.

Mortality and Transmitter Longevity

Of the 13 birds we implanted in 1999, two males and three females died within 14 days of surgery (38% mortality). A third male died at 16 days post-release. Finally, a fourth female died during spring migration near the outlet of Tetlin Lake (63.0°N, 142.7°W) approximately 35 days after capture. A subsistence hunting camp was present in this location and we suspect she was shot. We did not confirm mortality in any other birds. During July and August of 1999 we stopped receiving transmissions from two females (Fig. X). The fate of the birds could not be determined but the internal temperature sensor indicated the birds were alive at the time of the last transmission. The remaining transmitters failed between 23 November and 25 December 2000.

Of the 18 birds we implanted in 2000, two birds died in captivity (11%). A male died 9 days post-surgery and we euthanized a female 21 days post-surgery. Cause of mortality in the former was not determined. The latter likely resulted from a bacterial granuloma (introduced during surgery) with additional stresses from ureter trematodiasis (J. Raymond, DVM, NWZooPath, histopathology report). A third male died when attacked by a bald eagle (*Haliaeetus leucocephalus*) immediately after release. In July and August of 2000 we lost transmission from two males and one female (Figs. 4B, 5A, 6A). As with the 1999 birds, the temperature sensor indicated the birds were alive as of the last transmission. The remaining transmitters failed between 7 November 2000 and 31 March 2001. Individual transmitter performance varied by the quality and quantity of signals we received throughout the study.

DISCUSSION

Using satellite telemetry we described migration patterns and linked PWS wintering areas with breeding sites in the Northwestern Canada and molting sites in the Beaufort and Bering seas and GOA. This is the first successful attempt to use satellite telemetry to document movements of white-winged scoters. We compared movements of PWS birds with satellite transmitted white-winged scoters marked in wintering areas in 1) southeast Alaska (n = 2) (authors, unpublished data), 2) British Columbia (BC) (n = 23) (Sea duck Joint Venture 2006), 3) Washington State (WA) (n = 14) (D. Nysewander and J. Evenson, WDF&W, unpublished data), and 4) a breeding area on the Yukon Flats in interior Alaska (n = 7) (authors, unpublished data).

Sample sizes are small but precision is high and we felt it important to try and incorporate all this data recognizing the usual caveats associated with small sample sizes both spatially and temporally (Lindberg and Walker, in press). Comparisons with southeast and interior Alaska are especially provisional. However, scoters migrate in flocks of 10 to one hundred or more birds (Murie 1959, Irving 1960, Salter 1972, Mauer and Wilbor 1989) and each transmitted bird potentially represents many birds.

Spring Migration

Inland migration

White-winged scoters have often been observed during spring migration (Johnson and Herter 1989), but migration routes have not been described, as observers could not associate piecemeal observations with both origins and destinations. Isleib and Kessel (1973) observed white-winged scoters departing PWS in spring flying northward directly over mountain ranges and Irving (1960) observed migrants moving northward in the upper Copper River Valley and others following the Porcupine River to the Yukon Territory.

We described partial migration routes from PWS to nesting areas in Old Crow, YT and the NWT. While the information is incomplete, the southern portions of these routes (at minimum) are largely distinct from migrants from southeast Alaska (authors, unpublished data), and the Pacific coasts of BC (Sea duck Joint Venture 2006), and WA (D. Nysewander and J. Evenson, WDF&W, unpublished data). These latter birds all traveled inland east of Cape Spencer (58.2°N, 136.6°W) and over 600km east of PWS.

The timing of migration we observed was consistent with earlier observations in late May (Murie 1959, Lensink 1962, Campbell and Shepard 1973, Salter et al. 1974). We did not confirm spring staging areas, but our data suggest birds briefly staged near the confluence of the Black and Porcupine Rivers and between the Peel and Mackenzie Rivers in late May while waiting for more favorable weather conditions and advances in spring phenology.

Coastal migration

Several birds migrated directly to molting areas. These birds followed coastal routes, both east and west from PWS and generally departed later than inland migrants. Immature birds undergo pre-molt migrations directly from the winter area to the molting area (Salomonsen 1968) but little is known about older nonbreeding birds. Our sample of bursal measurements and number of coastal migrants was too small to draw conclusions about age class distribution between inland and coastal migrants but an ATY female and a TY male bypassed the breeding area.

The few coastal migrants made it difficult to identify patterns. Spring migration was not unidirectional. One male migrated east from PWS to molt near Cape Yakataga, a much shorter distance by about 80%, than the route to westerly molting areas in the southeastern Bering Sea. Some westerly migrants passed through lower Cook Inlet, possibly timed to coincide with late herring spawn events (Nature Conservancy 2003). Spring migration through this area has not been well documented but fisheries researchers have reported several thousand white-winged scoters plus other sea ducks during peak herring spawn events in mid May (Ted Otis, pers. comm.). One bird remained in Cook Inlet to molt while others proceeded to Bristol Bay likely taking overland routes across the Alaska Peninsula.

One nonbreeding male migrated west past Cape Pierce (58.8°N, 159.6°W) in western Alaska along a route where Herter et al. (1989) estimated 51,200 pre-molting white-winged scoters, primarily adult males. Two other PWS birds, a male and female remained southeast of Cape Pierce. All three birds arrived in the area past peak herring spawn in the Togiak district though possibly in time for late spawning events (ADF&G 2000). A distributional response of sea ducks to herring spawn has been established on wintering areas in PWS and BC (Bishop and Green 1981, Boyd, 2003) but the response among migrants is not well known.

We preferred to catch birds that strictly wintered in PWS rather than migrants. Isleib and Kessel (1973) observed numerous spring migrants along the outer coasts of PWS but did not mention an influx of birds staging in PWS. We have no record of migrants from southwest Alaska traveling through or staging in PWS in spring (authors unpublished data) consistent with observations of migrants from western Alaska traveling inland (Murie 1959). Nor did birds from BC, WA, or southeast Alaska migrate through PWS (see above). In fall, thousands of scoters stage along the coast to the southeast of PWS between Cape Suckling and Cape Yakataga (Conant and Groves 2001). If these birds remained there to winter they would provide a potential source of spring migrants. We transmitted one and possibly two birds that wintered in this area the following year and we suspect late winter or spring migrants from these nearby coastal wintering areas move to PWS and join wintering birds.

Breeding

Breeding Areas

PWS birds used three widely separated breeding areas within the boreal forests of western Canada and Alaska (Fig. 5C, 8), (Palmer 1976, Bellrose 1980) with the majority of breeding concentrated in the NWT (Fig. 8). These areas were nearly distinct from breeding areas used by birds from coastal BC (Sea Duck Joint Venture 2006) and WA (D. Nysewander and J. Evenson, WDF&W, unpublished data). The longitudinal divide, with some overlap occurred along the western arms of Great Bear Lake in the NWT (ca. 123.3°W). This marked the approximate western extent of the breeding range for BC and WA populations and the eastern extent for PWS birds. From here, the breeding range of BC and WA birds extended far to the south and east while PWS birds extended west. Birds from southeast and southwest Alaska extended into the range of PWS birds from the east and west respectively. Population affiliations may occur within regions (e.g. within the core breeding area) at smaller spatial scales than we were able to detect.

Relationship to North American Waterfowl Population Surveys

The majority of the core nesting area we identified in the NWT is surveyed annually as part of the Continental Waterfowl Breeding Population Habitat Survey (WBPHS) conducted each May since 1955 (Smith 1995). Most scoter population estimates are derived from this survey. The WBPHS divides the U.S. and Canada into over 50 habitat

strata. PWS birds used only three of these strata during the breeding season (from west to east with increasing in frequency): stratum 4 (Yukon Flats, Alaska), stratum 12 (Old Crow Flats, YT), and stratum 14 (NWT forest tundra), (Smith 1995). White-winged scoters are the most abundant scoter in each of these strata (Irving 1960, Lensink 1962, Haszard 2001). The Yukon Flats was used by only one male, possibly an anomaly for PWS birds, or at least not representative of breeding females and Old Crow was used by just two females. PWS birds were most represented in Stratum 14.

Stratum 14 encompasses 127,484 km² (Smith 1995) or about triple the size of the core breeding area we described for PWS and supports 30% of the surveyed continental scoter breeding population (S. Slattery, Ducks Unlimited, Canada, pers. comm.). Most of the core breeding area lies within the eastern third of stratum 14, while the most southerly and easterly portions lie outside the boundary of the WBPHS. In the western portion of stratum 14 PWS birds mixed with white-winged scoters from southeast and southwest Alaska (authors, unpublished data). Unfortunately we cannot use the WBPHS to assess changes in PWS populations without population assessment at smaller spatial scales and population surveys capable of identifying shorter-term trends in individual species. The many shortcomings of the WBPHS preclude good assessment of scoter population trends, especially in the short-term (Smith 1995, Alberta Sustainable Resource Development 2002). We have presented the beginnings of this long, multi-disciplined process and eventually we hope the results of population delineation studies will guide the timing and locations of surveys.

Molting

Several authors observed post-breeding movements or identified molting areas in coastal Canada and Alaska (Johnson and Herter 1989, Herter et al 1989, Alexander et al 1988, Salter et al. 1980, Johnson and Richardson 1982, Henny et al 1995, Mauer and Wilbor 1989), but this study is the first to link molting locations with specific breeding and wintering areas. We identified links to several coastal molting areas in northern Canada and western Alaska. Most conclusive among these is the affiliation, though not absolute, of PWS wintering birds to breeding and molting areas in the NWT. Molting areas used by PWS birds were apparently discrete from molting areas used by female white-winged scoters from BC and WA (Sea Duck Joint Venture 2006, D. Nysewander and J. Evenson, WDF&W, unpublished data). Wintering and breeding birds may have affinity to certain molting areas but individual molting areas likely support aggregates of flocks (scale of large bays, e.g. Liverpool Bay, Kuskokwim Bay) from several widely separated breeding and wintering areas. Fidelity to these sites as well as further structuring at smaller spatial scales is unknown.

Beaufort Sea

Only birds that occupied breeding areas in the NWT and Yukon Flats molted in the Beaufort Sea, primarily in Liverpool Bay (Fig. 9). Liverpool Bay contains 3 of 6 key scoter molting areas between Herschel Island, YT and Cape Bathurst, NWT (Barry 1976, Alexander et al. 1988, Dixon and Gilchrist 2002) and the number of molting sea ducks in

these areas is thought to have declined considerably since the early 1970's (Barry 1974, Alexander et al. 1988, Johnson and Herter 1989). Post-breeding birds originating from wintering areas in southwestern and southeast Alaska (authors, unpublished data) also molted in Liverpool Bay so it is not unique to PWS birds but we suspect a high proportion of PWS breeding birds used Liverpool Bay. One male each from BC and (Sea Duck Joint Venture 2006) and WA (D. Nysewander and J. Evenson, WDF&W, unpublished data) staged here but molting was not confirmed and these birds were atypical of other birds from these areas.

Bering Sea

Birds were widely dispersed from Kuskokwim Bay to Nelson Lagoon (56.2°N, 159.3°W) on the north side of the Alaska Peninsula indicating many suitable molt sites in this region for breeders and nonbreeders. Only one bird, a female, migrated here from the NWT, so any affinity with PWS breeding birds may be low. Most birds molted from the Nushagak Peninsula to Kuskokwim Bay. No fall survey data is available for this area. Few white-winged scoters molted north of Kuskokwim Bay (Dau 1987) so the destination of the 51,200 pre-molting white-winged scoters observed by Herter et al. (1989) remains a mystery. The area between the Nushagak Peninsula and Kuskokwim Bay is a likely priority for future surveys during summer and fall.

One bird molted and later staged on the north side of the Alaska Peninsula near Nelson Lagoon in autumn of 1999 (Fig. 3A). Fewer than 1,000 white-winged scoters including 285 in near Port Moller/Nelson Lagoon were counted along the north side of the Alaskan Peninsula in early October 1999 when this bird was present (Mallek and Dau 2000). Over twenty years earlier, Gill et al. (1981) observed 1,000 molting white-winged scoters just in the Port Moller/Nelson Lagoon area in early August; 2,000 birds in late September; and several thousand in late October (Byrd reported in Gill et al 1981).

Gulf of Alaska

From 1990–1992, at minimum several hundred molting white-winged scoters died at Cape Suckling and Cape Yakataga. No definitive causes for the mortality could be identified (Henny et al. 1995). During these investigations which continued into 1993, observers estimated maximum numbers of living scoters as follows: 1,000 at Cape Suckling, 1,870 between Cape Suckling and Cape Yakataga, and 1,200 at Cape Yakataga. Species were not differentiated but based on carcass counts white-winged scoters represented the large majority of these offshore observations. Investigators were frustrated by a lack of knowledge about migration patterns of these birds prior to their death (Henny et al. 1995). The movements we described for Cape Yakataga and Cape Suckling birds would have given investigators the benefit of a starting point. And will help researchers focus sampling and population monitoring efforts in the event of future die-offs.

Birds from other Pacific wintering areas did not molt in the northern GOA although two birds from WA passed through the area during migration, in June and August

respectively (D. Nysewander and J. Evenson, WDF&W, unpublished data). Additional surveys are needed to identify the importance of the various molt sites we have described. These sites likely support all three species of scoters so differentiation is necessary to assess relative abundance.

Winter Fidelity

We assumed birds returning to PWS in late fall and early winter remained until the following spring but at best, we received location data for about 11 months. Once a bird returned to PWS we have no record of it migrating elsewhere.

Winter site fidelity has been documented for members of several waterfowl taxa including sea ducks (see Anderson et al. 1992, Robertson and Cooke 1999) but not specifically for white-winged scoters. Site fidelity is most advantageous in stable environments (Esler et al. 2002); familiarity of local habitat conditions and availability of mates are other possible selective factors (Anderson et al. 1992, Robertson and Cooke 1999). Site fidelity can be an evolutionary precursor to the formation of unique population segments.

We believe the high return rate to PWS among breeding birds indicates winter site fidelity in white-winged scoters. Dispersal may be more common among unpaired males especially in birds, like scoters, with male biased sex ratios. Regardless, relative levels of philopatry and dispersal may vary among years, locations, age, sex, and breeding status and it is premature to infer at what spatial and temporal scale winter site fidelity operates in PWS.

Current marine boat surveys in PWS indicate large annual variations in winter populations (Sullivan et al. 2005). Whether these surveys reflect relatively low or variable rates of intra-annual winter site fidelity; a high and varying percentage of dispersal by nonbreeding birds or production by breeding birds; or poor design for sampling scoters is unknown. We need to better quantify interannual rates of winter site fidelity and dispersal by age and breeding status and understand winter movements and distribution before we can design optimal surveys or interpret existing surveys.

Mortality

The most serious problem we encountered in this study was the high rate of mortality that occurred within two weeks of surgery in 1999. By holding birds in captivity in 2000 we increased survival. In both years scoters were in good physical condition and responded well to surgical procedures (Dr. D. M. Mulcahy pers. comm.) but in 1999, post-release mortality was greater than other sea duck satellite implant studies using similar surgical procedures: spectacled eiders (Petersen et al. 1995); king eiders (Dickson et al. 1998); white-winged scoters (authors unpublished data); Barrow's goldeneyes (Robert et al. 2000); and harlequin ducks (Mulcahy and Esler 1999.).

The primary difference between this study and others (excluding harlequin ducks) is that we implanted birds in late winter in the GOA. Surgical implants for the first three studies (spectacled eiders, king eiders, white-winged scoters) were conducted on the breeding grounds. Behavioral differences may be attributable to increased survival of harlequin ducks which feed near shore and are comfortable on land (Robertson and Goudie 1999) and behavioral and geographic differences may be factors when comparing Barrow's Goldeneyes. Our mortality rates were more comparable to those reported for alcids (murres) implanted with satellite transmitters (Hatch et al. 2000). These birds were also captured and released in marine waters of Alaska and feed at deeper depths.

In simplest terms, the relatively high mortality of post-operative scoters primarily resulted from altered behavior after surgery which increased vulnerability to predators. In 1999, we observed post-operative birds isolated from flocks showing signs of lethargy, and appearing less capable of diving for food or responding to an avian threat. In 2000, we were able to examine birds post-surgery. We observed irritation at the incision and antenna exit site, seepage from incisions, and possible infections. Some captive birds failed to maintain waterproofing. Had these birds not been allowed to recover from surgery in a protected environment we would have expected higher mortality rates.

PWS is a predator rich environment with about 6,000 bald eagles (Bowman et al. 1997). In addition to witnessing a bald eagle capture a transmitted bird, eagles were frequently observed near the trap site disturbing flocks of sea ducks and we recovered several transmitters from under eagle roost trees although we do not know if these were predated or scavenged. With or without predators post-surgical complications developed but in 1999, birds had less opportunity to recover.

We implanted birds in late winter because we hoped to monitor birds for most of their annual cycle culminating in a return to wintering areas. Late winter may be a particularly stressful period in the annual cycle of sea ducks in northern latitudes. Limited daylight hours during winter and prolonged and severe storm events may limit winter foraging. If birds are energetically compromised and surgical effects constrain a bird from diving for food, then energetic processes may be further altered. Esler et al. (2000) found lower survival rates in harlequin ducks as winter progressed. Birds may have difficulty compensating for the additional stresses of capture and surgery at this time of year. Storms immediately before capture or soon after release seem to increase mortality rates. Behavioral and physiological changes prior to migration may also affect survival.

By holding birds in captivity pre- and post-surgery we allowed birds to first acclimate to captivity and then recover in a sheltered environment that improved survival in spite of additional stressors created by transporting and holding birds. However, it added considerable cost to the project. Since we conducted this study, technological improvements (battery life, power consumption, and programming) have extended transmitter life. These provide greater flexibility in project planning. We can now capture birds in any season and monitor them for one to two annual cycles. For studies focusing on wintering birds, implanting birds in autumn, soon after their return may increase survival rates. Once affiliations are established, winter objectives may be accomplished

by marking birds at other seasons. We had better success marking scoters at breeding areas where warmer air and water temperatures, fewer predators, easier access to food and better escape cover likely led to increased survival (authors unpublished data). Our small sample sizes and year and location variability preclude statistical comparisons.

CONCLUSIONS

Scoters are widely dispersed throughout remote areas and we had little knowledge of the migration patterns, affiliations between seasonal use areas, and fidelity to wintering, molting, and breeding areas. This made it extremely difficult to obtain and interpret information on population trends, productivity, survival, and harvest. This was the first project to successfully use satellite telemetry to identify the timing and routes of migration, affiliations between seasonal use areas and winter fidelity in white-winged scoters.

Migration Patterns and Seasonal Use Areas

Migration

Migration between seasonal use areas was generally rapid and direct occurring within a few days to a week or more. We described a general spring migration route from PWS to the Yukon Flats, Old Crow Flats, and NWT that filled in knowledge gaps and provided additional context to earlier observations. Refining these routes may provide additional opportunities to study migration and link information to PWS wintering populations. The timing of spring migration was consistent with historical accounts. Timing and routes of migration were different between breeding and nonbreeding birds.

Breeding

Most PWS birds (81%) used breeding areas in the NWT where numbers may have declined by as much as 56% from 1977 to 1996 (Dixon and Gilchrist 2002). Comprehensive surveys and research efforts conducted within the NWT core breeding area may serve to describe productivity and population trends for PWS birds and be a useful starting point to investigate population delineation at smaller spatial scales. As population affiliations are identified new survey designs (Mallek 2001) need to be incorporated that reflect the distribution of known population segments. If affiliations can be established, winter surveys in PWS may be advantageous for assessing abundance and trends (see Winter Fidelity below).

Molting

Molt sites of PWS scoters were widespread and primarily distinct from breeding areas. We identified coastal molting areas in the Beaufort Sea, the southern Bering Sea, and the northern GOA. Liverpool Bay in the Canadian portion of the Beaufort Sea, NWT was the primary molting site for breeding birds. We identified a link between a core breeding area in the NWT and molting areas in Liverpool Bay. Population structuring may occur at

smaller scales in specific bays (i.e. smaller bays within Liverpool Bay) but this has not been investigated.

Nonbreeding birds bypassed breeding areas and migrated along the coast to molting areas in the southern Bering Sea or northern GOA. Most of these areas have not been surveyed during the molt and we do not know the significance of many of these sites, the annual fidelity to these sites, or the strength of affiliations with PWS. We suspect molting sites are aggregates of birds from several breeding and wintering areas. We identified a link between molting birds near Capes Yakataga and Suckling in the northern GOA with staging and perhaps wintering areas in PWS.

Affiliations

We identified affiliations of white-winged scoters from PWS with breeding and molting areas in Alaska and western Canada. Identifying these affiliations is especially important in a species like white-winged scoters due to their high rates of female philopatry to natal areas (Brown and Brown 1981) and their low rates of annual recruitment (Brown and Fredrickson 1997).

Most significantly we identified an affiliation between PWS and a breeding area in the Mackenzie, Anderson, and Horton River basins in the NWT and a molting site in Liverpool Bay, NWT. This is the first time precise migratory connectivity and habitat affiliations have been documented for white-winged scoters. Breeding and molting areas were nearly distinct (i.e. small geographic and numerical overlap) from those used by birds wintering further south along the coasts of BC and WA. More evidence is needed before we can solidify the spatial and temporal scale of this affiliation, but this type of association will help researchers and managers investigate, assess, or identify future resource related concerns.

Philopatry

We documented strong evidence for winter site fidelity in breeding males and females. Birds using breeding and molting areas in the NWT had a high return rate to PWS the following winter. Nonbreeding birds had lower return rates and may represent dispersal. This is the first time winter site fidelity has been documented for white-winged scoters. The strength of philopatry to wintering and molting areas is unknown.

Management and Research Implications

Factors affecting distribution, and ultimately population dynamics, may occur anywhere within a bird's annual cycle and may result from a variety of complex and interacting events and processes. Testing hypotheses regarding the relationship between migratory connectivity and its behavioral and evolutionary effects requires knowledge of geographical patterns of movements of migratory birds (Webster et al. 2002). While there are several valuable methods to study migratory connectivity, satellite telemetry provides

the greatest precision to define migratory pathways and affiliations between breeding, molting, and wintering areas (Webster al 2002).

We have taken the first step in this lengthy process by providing key focal points for research and management studies directed at declines in PWS populations. The mechanisms causing population change are poorly understood. Many causes for past population declines and impediments to future growth or stability have been speculated for the geographical areas we affiliated with PWS birds. These range from changes at global to local levels and all are difficult to test and quantify. For coastal areas these include: changes in ocean temperatures (Piatt and Anderson 1996, Grebmeier et al. 2006), coastal development (Dixon and Gilchrist 2002), exposure to contaminants including oil spills (Vermeer and Peakall 1979, Piatt et al. 1990, Henny et al. 1991, Olendorf et al. 1991, Henny et al. 1995), declines in herring stocks (Sharp et al. 2000), and increased competition from sea otters (Stratton 1981), while climate change and industrial development (Cohen 1997, Rouse et al. 1997, Mackenzie River Basin Board 2004) are leading candidates for altering habitats and affecting wildlife population dynamics in the NWT. If we are to narrow the range of possibilities we need to identify affiliations and focus local research projects and monitoring efforts in areas with known connectivity.

Our efforts establish a basis for future population delineation and improved monitoring efforts. We present this as a small step toward identifying the mechanisms of population change. We also hope it will be useful to regional planners. Much additional effort among many disciplines is required before we can draw conclusions about population structuring at both smaller and larger spatial scales. However, we emphasize the importance of further defining and refining seasonal affiliations so we can comprehensively assess mortality factors throughout the annual cycle.

ACKNOWLEDGEMENTS

We are greatly indebted to the many people who have contributed their time, energy, skills, and knowledge to the many aspects of this project. All have contributed to its success. Dr. Dan Mulcahy's and Dr. Pam Tuomi's skills, experience, and dedication were invaluable in the care, handling, and surgical implantation of satellite transmitters in our scoters. Celia Hall and Millie Gray provided an equally invaluable service to our project with their skills and patience as anesthetists. We thank Dave Crowley, Tom Rothe, Tim Bowman, Sam Iverson, John Ashley, and Joshua Hall for making the capture efforts go smoothly, safely, and enjoyably.

We gratefully acknowledge Susan Inglis and the staff at the Alaska SeaLife Center in Seward, Alaska for their time and dedication in the care and handling of captive scoters. We thank Dan Esler and Danielle Mather for their assistance with aging birds. We thank Dave Nysewander and Joe Evenson, Washington Department of Fish and Wildlife for sharing their data on scoter movements and Dave Douglas for providing ArcView filtering software. We are grateful to Captain Dean Rand and the crew of the MV Discovery, including Ken Hadzima and Heather and Hanna Rand for their patience and

hospitality as well as being ready, willing, and able to accommodate any needs that arose. The U.S. Coast Guard, Kodiak Operations and Scheduling Center provided logistical support in our efforts to retrieve satellite transmitters. We thank Gail and Steve Ranney for flying birds to and from the ASLC.

We thank Joshua Hall, and Jenifer Childress of the Chugach School District for producing the Youth Area Watch component of this project.

We extend thanks to Celia Rozen, ADF&G librarian, for her dedication and diligence in conducting literature searches, acquisitions, and help with editing along with and her many other library skills. Tom Rothe reviewed the draft manuscript. We also thank Bill Hauser, Melanie Bosch, and Cathy Kane for their support. And finally, we greatly appreciate the support of Molly McCammon, Stan Senner and the staff of the *Exxon Valdez* Oil Spill Restoration Office and the *Exxon Valdez* Oil Spill Trustee Council in funding this project.

LITERATURE CITED

- Alaska Department of Fish and Game. 2000. Division of Commercial Fisheries annual management report 2000 Bristol Bay Area. Regional Information Report1 No. 2A01-10. Anchorage. 160pp.
- Alberta Sustainable Resource Development. 2002. Status of white-winged scoter (*Melanitta fusca deglandi*) in Alberta. Alberta Sustainable Resource Development, Fish and Wildlife Division, and Alberta Conservation Association, Wildlife Status Report No. 45, Edmonton, AB. 15pp.
- Alexander, S.A., T.W. Barry, D.L. Dickson, H.D. Prus, and K.E. Smyth. 1988. Key areas for birds in coastal regions of the Canadian Beaufort Sea. Edmonton: Canadian Wildlife Service.
- Alisauskas, R. T., J.J. Traylor, C.J. Swoboda, and F.P. Kehoe. 2004. Components of population growth rate for White-winged Scoters in Saskatchewan, Canada. *Animal Biodiversity and Conservation*, 27.1: 451–460.
- Anderson, M.G., J. M. Rhymer, and F. C. Rohwer. 1992. Philopatry, dispersal, and the genetic structure of waterfowl populations, p. 365–395. In B. D. J. Batt, A. D. Afton, M. G. Anderson, C. D. Ankney, D. H. Johnson, J. A. Kadlec, and G. L. Krapu [EDS.], *Ecology and management of breeding waterfowl*. University of Minnesota Press, Minneapolis, MN.
- Arner, B.D., D.L. Dickson, and G. Verreault. 1985. Bird observations from Atkinson Point, Northwest Territories, 1984. *Can. Wild. Serv.*, Edmonton, Alberta. 37 pp.
- Bellrose, F.C. 1976. *Ducks, Geese, and Swans of North America*. Wildlife Management Institute. Stackpole Books, Harrisburg, Pa.

- Birket-Smith, K. 1953. The Chugach Eskimo. Nationalmuseets Skifter, etnografisk roekke, v. 6. Copenhagen.
- Bishop, M. A., and S.P. Green. 2001. Predation on Pacific herring (*Clupea pallasii*) spawn by birds in Prince William Sound, Alaska. Fisheries Oceanography 10 (suppl.1):149 – 158.
- Bowman, T.D, P.F. Schempf, and J.I. Hodges. 1997. Bald eagle population in Prince William Sound after the Exxon Valdez oil spill. Journal of Wildlife Management, 61(3): 962-967.
- Boyd, S. 2003. Numerical and distributional response of scoters to herring spawn events in the Strait of Georgia, B.C. Abstract. Third NA Duck symposium. Sacramento, CA.
- Brown, P.W., and M.A. Brown. 1981. Nesting biology of the White-winged Scoter. Journal of Wildlife Management 45:38-45.
- Brown, P. W. and L. H. Fredrickson. 1997. White-winged scoter (*Melanitta fusca*). In A. Poole and F. Gill, editors The birds of North America, No. 274. Academy of Natural Sciences, Philadelphia, Pennsylvania, and American Ornithologists' Union, Washington, D.C. 28 pp.
- Campbell, R.W. and M.G. Shepard. 1973. Spring waterfowl migration on the Mackenzie River from Norman Wells to the Arctic Red River, NWT, 1972. In: Towards an environmental impact assessment of the portion of the Mackenzie Valley gas pipeline from Alaska to Alberta. Envir. Protection Board, Winnipeg.
- Canadian Wildlife Service Waterfowl Committee. 2000. Population status of migratory game birds in Canada: November 2000. CWS Migratory Birds Regul. Rep. No. 1.
- Cohen, S.J. 1997. What if and so what in northwest Canada: could climate change make a difference to the future of the Mackenzie Basin? Arctic 50(4):293-307.
- Conant, B. and D. Groves. 2001. Gulf of Alaska shoreline waterbird surveys 2001. Unpublished rept. U.S. Fish and Wildlife Service, Juneau. 10pp.
- Dau, C.P. 1987. Birds in nearshore waters of the Yukon-Kuskokwim Delta, Alaska. The Murrelet 68(1):12-23.
- Dickson, D.L. and H.G. Gilchrist. 2002. Status of marine birds of the southeastern Beaufort Sea. Arctic 55(1): 46-58.
- Dickson, D.L., R.S. Suydam, and J. Fadely. 1998. Tracking the movement of king eiders from nesting grounds in western arctic Canada to their molting and wintering areas using satellite telemetry. 1997/98 Progress Rept. Canadian Wildl. Serv., Edmonton, AB. 7pp + tabs, figs, append.

- Ely, C.R., D.C. Douglas, A.C. Fowler, C.A. Babcock, and D.V. Derksen. 1997. Migration behavior of tundra swans from the Yukon-Kuskokwim Delta, Alaska. *Wilson Bull.* 109(4):679-692.
- Esler, D. 2000. Applying metapopulation theory to conservation of migratory birds. *Conservation Biology* 14:366-372.
- Esler, D., J. A. Schmutz, R. L. Jarvis, and D. M. Mulcahy. 2000. Winter survival of adult female harlequin ducks in relation to history of contamination by the Exxon Valdez oil spill. *Journal of Wildlife Management* 64:839-847.
- Esler, D., T. D. Bowman, K. A. Trust, B. E. Ballachey, T. A. Dean, S. C. Jewett, and C. E. O'Clair. 2002. Harlequin duck population recovery following the 'Exxon Valdez' oil spill: progress, process and constraints. *Mar. Ecol. Prog. Ser.* 241:271-286.
- Fancy, S.G., L. F. Pank, D. C. Douglas, C. H. Curby, G. W. Garner, S. C. Amstrup, and W. L. Regelin. 1988. Satellite telemetry: a new tool for wildlife research and management. *U.S. Fish Wildl. Serv., Resour. Publ.* 172pp.
- Gabrielson I.N. and F.C. Lincoln, 1959. *Birds Of Alaska*. Stackpole Books, Harrisburg, Pa, and The Wildl. Manage. Institute, Wash. D.C.
- Gill, R.E., M.R. Petersen, and P.D. Jorgensen. 1981. Birds of the Northcentral Alaska Peninsula, 1976-1980. *Arctic* 34(4):286-306.
- Godfrey, W.E. 1986. *The Birds of Canada*. National Mus. of Nat'l Sciences. Ottawa. 595pp.
- Goudie, R.I., S. Brault, B. Conant, A. V. Kondratyev, M. R. Petersen, and K. Vermeer. 1994. The status of sea ducks in the North Pacific Rim: Toward their conservation and management. *Trans. 59th N. Amer. Wildl. Natur. Resour. Conf.*:27-49.
- Grebmeier, J.M., J.E. Overland, S.E. Moore, E.V. Farley, E.C. Carmack, L.W. Cooper, K.E. Frey, J.H. Helle, F.A. McLaughlin, S. L. McNutt. 2006. A major ecosystem shift in the northern Bering Sea. *Science* 311:1461-1464.
- Gustafson, M. E., J. Hildenbrand and L. Metras. 1997. *The North American Bird Banding Manual (Electronic Version)*. Version 1.0
- Gwich'in Elders. 1997. Nahn' Kak Geenjit Gwich'in Ginjik: Gwich'in words about the land. Inuvik, Northwest Territories: Gwich'in Renewable Resources Board.
- Haszard, S.L. 2001. Habitat requirements of white-winged and surf scoters in the Mackenzie Delta region, Northwest Territories. *Arctic* 54:472-473.

- Hatch S.A., P.M. Meyers, D.M. Mulcahy, and D.C. Douglas. 2000. Performance of implantable satellite transmitters in diving seabirds. *Waterbirds*. 23(1):84-94.
- Henny, C.J., L.J. Blus, R.A. Grove, and S.P. Thompson. 1991. Accumulation of trace elements and organochlorines by surf scoters wintering in the Pacific Northwest. *Northwest Naturalist*. 72:43-60.
- Henny, C.J., D.B. Rudis, T.J. Roffe, and E. Robinson-Wilson. 1995. Contaminants and sea ducks in Alaska and the circumpolar region. *Environ. Health Perspec.* 103(Suppl 4):41-49.
- Herter, D.R., S.M. Johnston, and A.P. Woodman. 1989. Molt migration of scoters at Cape Pierce, Alaska. *Arctic*. 42(3): 248-252.
- Hodges, J.I., King J.G., Conant, B., and H.A. Hanson. 1996. Aerial surveys of waterbirds in Alaska 1957-94: Population trends and observer variability. Information and Technology Report 4. USDI, Nat'l Biological Service U.S. Fish and Wildlife Service, Juneau.
- Houston, C.S. and P.W. Brown. 1983. Recoveries of Saskatchewan-banded white-winged scoters, *Melanitta fusca*. *Can. Field-Natur.* 97:454-455.
- Huntington, H.P., R.S. Suydam, and D.H. Rosenberg. 2004. Traditional knowledge and satellite tracking as complimentary approaches to ecological understanding. *Environmental Conserv.* 31(3): 177-180.
- Isleib, M.E. and B. Kessel. 1973. Birds of the North Gulf Coast-Prince William Sound, Alaska. *Biol. Pap. Univ. Alaska* No. 14. 149 pp.
- Irving, L. 1960. Birds of Anaktuvuk Pass, Kobuk, and Old Crow. a study in Arctic adaptation. U.S. National Museum. Bulletin 217. Smithsonian Institution, Washington D.C. 409pp.
- Johnson, S.R., and D.R. Herter. 1989. The birds of the Beaufort Sea. Anchorage, Alaska: BP Exploration (Alaska) Inc.
- Johnson, S.R., and W.J. Richardson. 1982. Waterbird migration near the Yukon and Alaskan coast of the Beaufort Sea:II. Moulting migration of seaducks in summer. *Arctic* 35(2): 291-301.
- Kaiser, G.W., A.E. Derocher, S. Grawford, M.J. Gill and I.A. Manley. 1995. A capture technique for marbled murrelets in coastal inlets. *Journal of Field Ornithology* 66:321-333.
- Kistchinski, A.A. 1973. Waterfowl in north-east Asia. *Wildfowl* 24:88-102.

- Klosiewski, S.P., and K.K. Laing. 1994. Marine bird populations of Prince William Sound, Alaska, before and after the Exxon Valdez oil spill. EVOS Bird Study No. 2. U.S. Fish and Wildl. Serv., Anchorage. 85pp.
- Korschgen, C.E., K.P. Kenow, A. Gendron-Fitzpatrick, W.L. Green, and F.J. Dein. 1996. Implanting intra-abdominal radiotransmitters with external whip antennas in ducks. J. Wildl. Manage. 60:132-137.
- Lance, B. K., D.B. Irons, S. J. Kendall, and L. L. McDonald. 2001. An evaluation of marine bird population trends following the Exxon Valdez oil spill, Prince William Sound, Alaska. Marine Pollution Bulletin 42: 298-309.
- Lensink, C.J. 1962. Waterfowl populations and production within the impoundment area of the proposed dam at Rampart on the Yukon River, Alaska. Unpubl. progress rept. Bureau of Sport Fisheries and Wildlife, Laurel MD. 66pp.
- Lindberg, M.S. and J. Walker. In press. Satellite telemetry in avian research: sample size considerations. J. of Wildl. Manage.
- Mackenzie River Basin Board. 2004. Mackenzie River Basin, state of the aquatic ecosystem report. Fort Smith, NT.
- Mallek, E. J. and C. P. Dau. 2000. Aerial Survey of Emperor Geese and Other Waterbirds in Southwestern Alaska, Fall 1999. U.S. Fish and Wildlife Service, Waterfowl Management. Unpublished report. Fairbanks. 19pp.
- Mallek, E. J. 2001. Replicate aerial scoter surveys of the Yukon Flats, Alaska- 2001. Unpubl. Rept. U.S. Fish and Wildlife Service. Fairbanks, AK 15pp.
- Mather, D.D. and D. Esler. 1999. Evaluation of bursal depth as an indicator of age class of harlequin ducks. Journal of field Ornithology 70:200-205.
- Mauer, F.J. and S. Wilbor. 1989. Raptor surveys along the Porcupine River, Alaska, 1989. ANWR Progress Report No. FY89-01. US Fish and Wildlife Service. Fairbanks. 2pp.
- Moritz, C. 1994. Defining 'evolutionarily significant units' for conservation. Trends in Ecology and Evolution 9:373-375.
- Mulcahy, D.M. and D. Esler. 1999. Surgical and immediate postrelease mortality of harlequin ducks (*Histrionicus histrionicus*) implanted with abdominal radio transmitters with percutaneous antennae. J. of Zoo and Wildl. Med. 30(3):397-401.
- Murie, O.J. 1959. Fauna of Aleutians and Alaska Peninsula. North American Fauna No. 61. US Fish and Wildlife Service, Washington DC

- Ohlendorf, H.M., K.C. Marois, R.W. Lowe, T.E. Harvey, and P.R. Kelly. 1991. Trace elements and organochlorines in surf scoters from San Francisco Bay, 1985. *Environ. Monit. Assess* 18:105-122.
- Ornithological Council. 1997. Guidelines to the use of wild birds in research. Internet version. (www.nmnh.si.edu/BIRDNET/GuideToUse). Current at April 4, 1999. Wash. D.C.
- Palmer R.S. 1976. Handbook of North American birds. Vol. 3. Yale Univ. Press. New Haven. 560pp.
- Petersen, M. R., D. C. Douglas, and D. M. Mulcahy. 1995. Use of implanted satellite transmitters to locate spectacled eiders at-sea. *Condor* 97:276-278.
- Piatt, J.F., C.J. Lensink, W. Butler, M. Kendziorek, and D.R. Nysewander. 1990. Immediate impact of the Exxon Valdez oil on marine birds. *The Auk* 107: 387-397.
- Piatt, J.F. and P. Anderson. 1996. Response of common murrelets to the Exxon Valdez oil spill and long-term changes in the Gulf of Alaska marine ecosystem. In: Rice, S.D., R.B. Spies, D.A. Wolfe and B.A. Wright, editors. Exxon Valdez Oil Spill Symposium Proceedings. American Fisheries Society Symposium 18:720-737.
- Porslid, A.E. 1943. Birds of the Mackenzie Delta. *Can. Field Nat.* 54(2):19-35.
- Powell, B.H. 2000. Game bird encounters and summarized bandings thru August 2000. Compact Disk. U.S. Geological Resources, Biological Resources Division. Laurel, MD.
- Robert, M., D. Bordage, J-P. L. Savard, G. Fitzgerald, and F. Morneau. 2000. The breeding range of the Barrow's Goldeneye in eastern North America. *Wilson Bull.* 112(1):1-7.
- Rouse, W.R., M. S. V. Douglas, R.E. Hecky, A. E. Hershey, G. W. Kling, L. Lesack, P. Marsh, M. McDonald, B. J. Nicholson, N. T. Roulet, J. P. Smol. 1997. Effects of climate change on the freshwaters of arctic and subarctic North America. *Hydrological Processes* 11 (8): 873-902.
- Rosenberg, D.H. and M.P. Petrula 1999. Scoter life history and ecology: Linking satellite technology with traditional knowledge, Exxon Valdez Oil Spill Restoration Proj. Ann. Rept. (Restor. Proj. 98273), Alaska Dept. of Fish and Game, Division of Wildl. Conser., Anchorage. 34pp.
- Rosenberg, D.H. and M.P. Petrula 2000. Scoter life history and ecology: Linking satellite technology with traditional knowledge, Exxon Valdez Oil Spill Restoration Proj. Ann. Rept. (Restor. Proj. 99273), Alaska Dept. of Fish and Game, Division of Wildl. Conser., Anchorage. 33pp.

- Robertson, G. J., and F. Cooke. 1999. Winter philopatry in migratory waterfowl. *Auk* 116:20–34.
- Robertson, G. J., and R. I. Goudie. 1999. Harlequin duck. *The birds of North America*. The American Ornithologists Union, Washington, D. C., USA, and The Academy of Natural Sciences, Philadelphia, Pennsylvania, USA.
- Salomonsen, F. 1968. The moult migration. *Wildfowl* 19:5-24.
- Safine, D. E. 2005. Breeding ecology of white-winged scoters on the Yukon Flats, Alaska. MscThesis. Univ. of Alaska Fairbanks. 114pp.
- Salter, R.E. 1972. Bird migrations on the north slope and in the Mackenzie River Valley regions, 1972. *Arctic Gas Biol. Rept. Series V. 13*. Ed: W.W.H. Gunn and J.L. Livingston. 161pp.
- Salter, R.E., W.J. Richardson, and C. Holdsworth 1974. Spring migration of birds through the Mackenzie Valley, NWT, April-May, 1973. *Arctic Gas Biol. Rept. Ser.* 28:1-160.
- Salter, R.E., M.A. Gallop, S.R. Johnson, and W.R. Koski, 1980. Distribution and abundance of birds on the Arctic Coastal Plain of Northern Yukon and adjacent Northwest Territories, 1971-1976. *Can. Field Nat.* 94 (3): 219-238.
- Schweigert, J., F. Funk, K. Oda, and T. Moore. 2002. Herring size-at-age variation in the North Pacific. In: *PICES-GLOBEC International Program on Climate Change and Carrying Capacity, Report of 2001 BASS/MODEL, MONITOR, and REX Workshops, and the 2002 MODEL Workshop*. Report no. 20. North Pacific Marine Science Organization, Sydney, BC, Canada.
- Scott, C.L., A.W. Paige, G. Jennings, and L. Brown. 1996. *Community Profile Database Catalog, Vol. 2*. Alaska Department of Fish and Game, Div. of Subsistence. Juneau.
- Sea Duck Joint Venture Management Board. 2001. *Sea Duck Joint Venture Strategic Plan: 2001–2006*. SDJV Continental Technical Team. Unpubl. Rept.[c/o USFWS, Anchorage, AK; Sackville, NB. 14pp. + append.
- Sea Duck Joint Venture. 2006. Satellite telemetry studies, white-winged scoters marked in British Columbia [maps]. http://www.seaduckjv.org/sts_bc_maps.html. 2006.
- Service Argos. 1996. *Users Manual*. Service Argos, Inc., CLS. Landover, MD. 176pp.
- Smith, G. W. 1995. A Critical review of the aerial and ground surveys of breeding waterfowl in North America. *Biological Science Report 5*, July 1995. National Biological Service, Technical Report Series. Washington, DC.

- Stratton, L. 1990. Resource harvest and use in Tatitlek, Alaska. Tech. Paper No. 181. Alaska Department of Fish and Game, Div. of Subsistence. Anchorage. 163 pp.
- Stephensen, S.W., D. B. Irons, S. J. Kendall, B. K. Lance, and L. L. McDonald. 2001. Marine Bird and Sea Otter Population Abundance of Prince William Sound, Alaska: Trends following the *T/V Exxon Valdez* Oil Spill, 1989-2000. Exxon Valdez Oil Spill Restoration Project Annual Report (Restoration Project 00159), U. S. Fish and Wildlife Service, Anchorage, Alaska. 115pp.
- Stratton, L. and E. B. Chisum. 1986. Resource use patterns in Chenega, western Prince William Sound: Chenega in the 1960's and Chenega Bay 1984-1986. Alaska Department of Fish and Game, Div. of Subsistence. Tech. Paper no. 139. Anchorage. 161 pp.
- Sullivan, K.M., A.E. McKnight, D.B. Irons, S.W. Stephensen, and S. Howlin. 2005. Marine bird and sea otter population abundance of Prince William Sound, Alaska: trends following the T/V Exxon Valdez oil spill, 1989-2004. Exxon Valdez Oil Spill Restoration Project Annual Report (Restoration Project 04159), U. S. Fish and Wildlife Service, Anchorage, Alaska.
- U.S. Fish and Wildlife Service. 1999. Population status and trends of sea ducks in Alaska. Migratory Bird Manage. Anchorage. 137 pp.
- Vermeer, K, and D.B. Peakall. 1979. Trace metals in seaducks of the Fraser River Delta intertidal area, British Columbia. Marine Pollution Bulletin. Vol. 10. pp189-193.
- Wahrhaftig, C. 1965. Physiographic divisions of Alaska. U.S. Geological Survey Professional Paper No. 482. U.S. Government Printing Office, Washington, D.C.
- Webster, M.S., P.P. Marra, S.M. Haig, S. Bensch, and R.T. Holmes. 2002. Links between worlds: unraveling migratory connectivity. Trends in Ecology and Evolution 17(2): 76-83.
- Witten, E. 2003. Tuxedni Bay to Kamishak Bay, Cook Inlet: Alaska Conservation Plan. Alaska Natural Heritage Program, Environment and Natural Resources Institute University of Alaska Anchorage. 80pp.
- Wolfe, R.J., A.W. Paige, and C.L. Scott. 1990. The subsistence harvest of migratory birds in Alaska. Alaska Department of Fish and Game, Division of Subsistence, Tech. Paper No. 197. Anchorage 183 pp.

Table 1. Summary of year (1=1999, 2=2000), sex (F= female, M=male), capture location (CL); and general location of breeding, molting, and wintering areas; and last location and date we received data for individual white-winged scoters captured in Prince William Sound, Alaska (PWS) and implanted with satellite transmitters in 1999 and 2000.

Year	Bird ID	Sex	CL ¹	Breeding Area ²	Molting Area	Winter Area	Last Location	Last Date (M/D/Y)
1	20375	F	SMB	Carcajou Lake, NWT	NA ³	NA	Breeding Area	8/8/99
1	23887	F	SMB	Anderson River, NWT	Port Moller, AK	Port Moller, AK ⁴	Port Moller, AK	11/28/99
1	23890	F	SMB	Anderson River, NWT	Liverpool Bay, NWT	Port Fidalgo, PWS	PWS	12/25/99
1	23894	F	SMB	Mackenzie River, NWT	Liverpool Bay, NWT ⁴	NA	Liverpool Bay NWT	7/24/99
1	20376	M	SMB	Nonbreeder	Cook Inlet, AK	Chiginagak Bay, AK	Chiginagak Bay, AK	12/22/99
1	24124	M	SMB	Nonbreeder	Togiak Bay, AK	Naked Island, PWS	PWS	11/23/99
2	20377	F	OI	Old Crow, YT	Knowles Bay, PWS	Sheep Pt., PWS	PWS	1/17/01
2	23889	F	OI	Lac Maunoir, NWT	Liverpool Bay, NWT	Naked Island, PWS	PWS	2/2/01
2	27596	F	OI	Tadek Lake, NWT	NA	NA	Tadek Lake, NWT	7/9/00
2	27597	F	OI	Great Bear Lake, NWT	Liverpool Bay, NWT	Simpson Bay, PWS	PWS	1/16/01
2	27599	F	OI	Nonbreeder	Nushagak Peninsula, AK	Bristol Bay, AK	Becharof Lake, AK	3/15/01
2	30046	F	OI	Great Bear Lake, NWT	Great Bear Lake, NWT	Glacier Isl., PWS	PWS	12/22/00
2	30048	F	OI	Tadenet Lake, NWT	Tadenet Lake, NWT	Green Isl., PWS	PWS	1/14/01
2	30049	F	OI	Old Crow, YT	Kuskokwim Bay, AK	Umnak Isl., AK	Umnak Isl., AK	1/28/01
2	12992	M	OI	Lac Maunoir, NWT	Liverpool Bay, NWT ⁴	NA	Liverpool Bay, NWT	7/6/00
2	23888	M	OI	Yukon Flats, AK	Liverpool Bay, NWT	Knowles Bay, PWS	PWS	11/8/00
2	29992	M	OI	Great Bear Lake, NWT	Liverpool Bay, NWT	Glacier Isl., PWS	PWS	3/31/01
2	29994	M	OI	Great Bear Lake, NWT	Franklin Bay, NWT	Smith Isl., PWS	PWS	1/2/01
2	29995	M	OI	Nonbreeder	Cape Yakataga, AK	Icy Bay, AK	Icy Bay, AK	2/17/01
2	30047	M	OI	Great Bear Lake, NWT	Cape Suckling ⁴	NA	Kayak Island, AK	8/27/00
2	30051	M	OI	Nonbreeder	Cape Newenham, AK	NA	Cape Newenham, AK	11/29/00

¹ St. Matthews Bay, PWS (SMB), Orca Inlet, PWS (OI)

² Northwest Territories, Canada (NWT), Yukon Territory, Canada (YT)

³ Data not available (NA)

⁴ Presumed affiliation. Transmitter failed prior to verification.

Table 2. Median departure date (range) from coastal capture areas in Prince William Sound (PWS) (1999, 2000) by reproductive status (RS) and median arrival dates (range) and duration by sex (Female/Male) at breeding areas. Mean arrival dates for nonbreeding birds (NB) going directly to coastal molting areas are not included (refer to Table 3).

Year	RS ¹	PWS Capture Location	#F/#M ²	Winter Area Median Departure Date (Range)	#F/#M ²	Breeding Area Median Arrival Date (Range)	#F/#M ²	Breeding Area Mean No. Days (Range)
1999	B	St. Matthew's Bay, 60.8°N, 146.3°W	5/0	31 May (27 May – 3 Jun)	3/0	6 Jun (3–13 Jun)	3/0	51 (34-72) ³
	NB		0/2	1 June (31 May – 2 Jun)	0/2	—	—	—
2000	B	Orca Inlet 60.5°N, 145.9°W	7/5	24 May (18–27 May)	7/5	31 May (25 May – 8 Jun)	5/3	66 (15-107)/27 (16-44)
	NB		1/2	25 June (14 Jun – 8 Jul)	1/2	—	—	—

¹ Migrated to inland breeding area (B) or remained on marine waters (NB) after departing wintering area.

² F (Female), M (Male).

³ A fourth female spent 62 days at a breeding area before we lost transmission.

Table 3. Median arrival and departure dates and duration at molting areas and arrival dates at wintering areas by sex (Female/Male) and reproductive status (RS) for birds captured in late-winter in Prince William Sound (1999, 2000). Birds that bred and molted or molted and wintered in the same area are excluded from molting and winter arrival date and duration calculations.

RS	Year	Molting Area					Winter Area	
		F/M	Arrival	F/M	Departure	Mean # days (range)	F/M	Arrival ¹
Breeding	1999	3/0	25 Jul (9 Jul-23 Aug)	1/0	4 Oct	69	1/0	7 Oct
	2000	4/4	20 Jul (14 Jul – 9 Aug)/ 30 Jun (21 Jun- 25 Jul)	3/1	29 Sep (15 Sep – 28 Nov)/ 28 Sep ²	87 ³ (75-111)/ 95	5/3	25 Sep (15 Sep – 1 Dec) ⁴ / 1 Oct (21 Sep- 11 Oct)
Non-breeding	1999	0/2	16 Jun (13-20 Jun)	0/2	30 Aug (29 Aug – 1 Sep)	74 (72-76)	0/2	4 Sep (3 Sep – 4 Sep)
	2000	1/2	NA ⁵ /20 Jun (14 – 25 Jun)	1/2	11 Oct/4 Sep (28 Aug – 12 Sep)	72 ⁶ /77 (75-79)	1/1	11 Oct/29 Aug

¹ Does not include 2 females that molted at breeding area.

² Poor location data before and after departing molting area allowed calculations for only one male.

³ Does not include birds that molted and wintered in same area.

⁴ Two birds from returning from the Northwest Territories were estimated to return to the northern Gulf of Alaska by 10 July (male) and 27 July (female) to molt and winter. The male is not included in the table due to poor transmitter performance.

⁵ A non-breeding female was present in Bristol Bay on 31 July, the first location data since 16 July.

⁶ Minimum estimate.



Figure 1. Map of Prince William Sound, Alaska showing capture locations in 1999 and 2000 for white-winged scoters implanted with satellite transmitters.

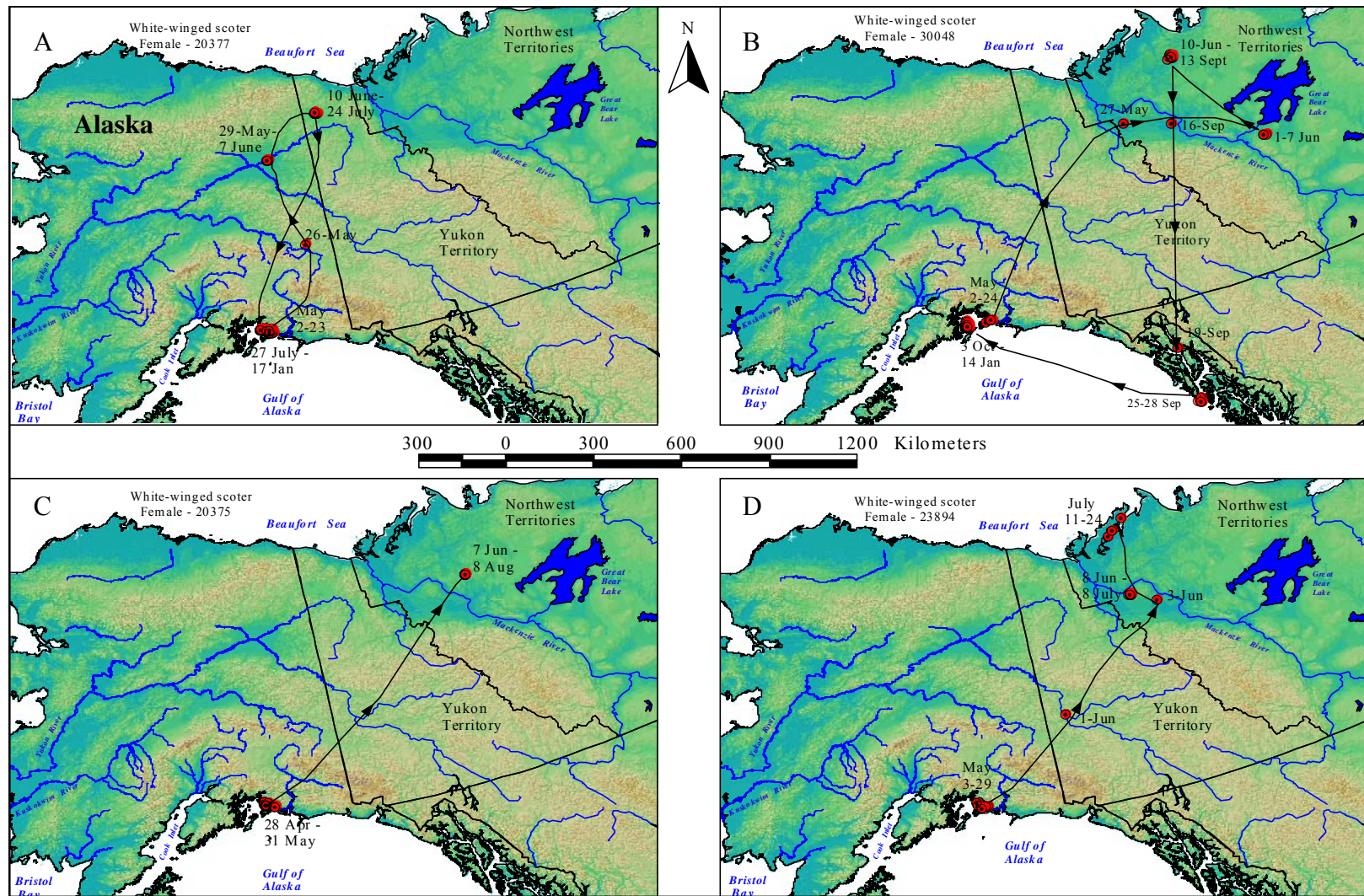


Figure 2. Movements of female white-winged scoters 20377, 30048, 20375, and 23894 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes.

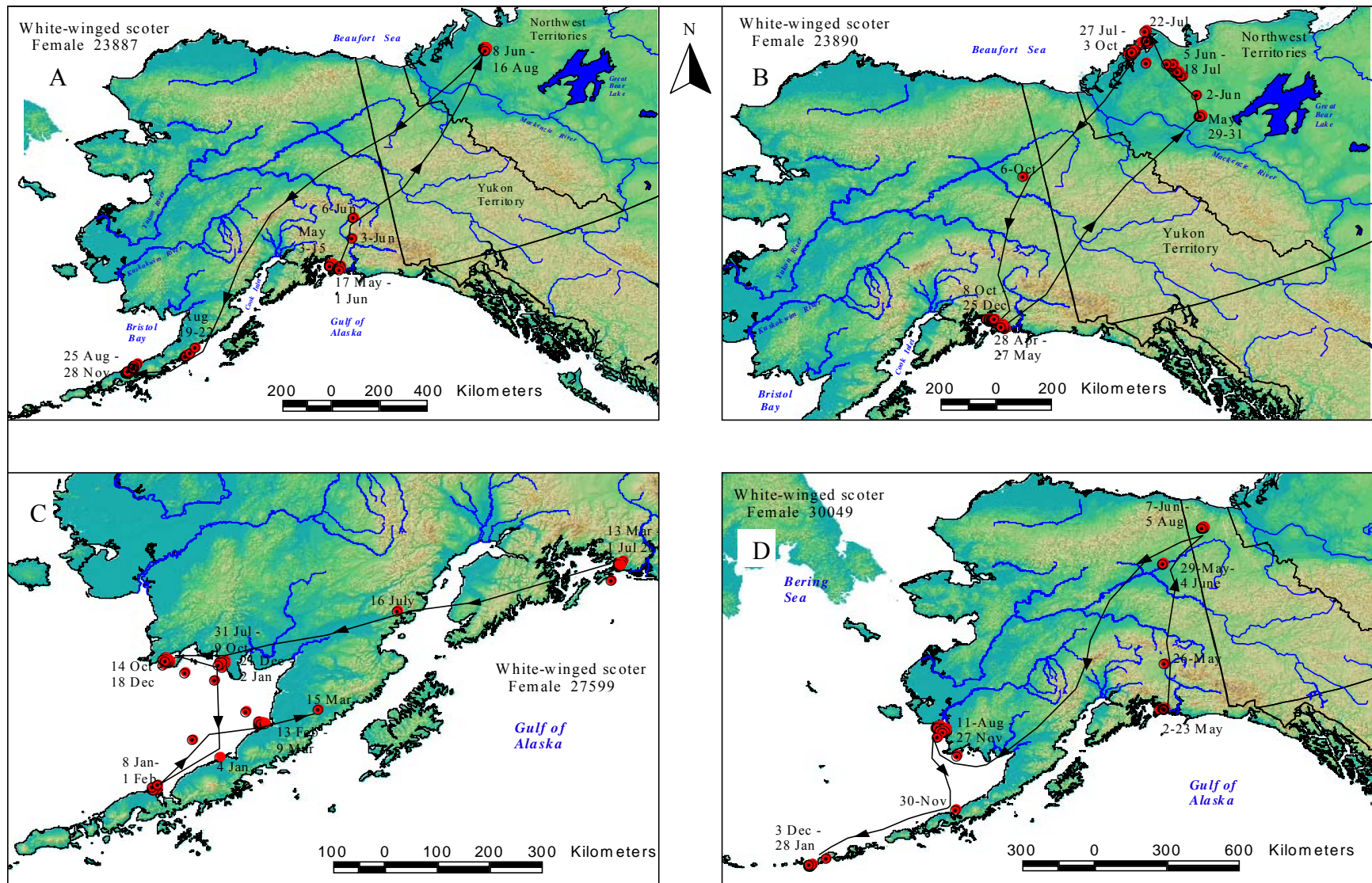


Figure 3. Movements of female white-winged scoters 23887, 23890, 27599, and 30049 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes.

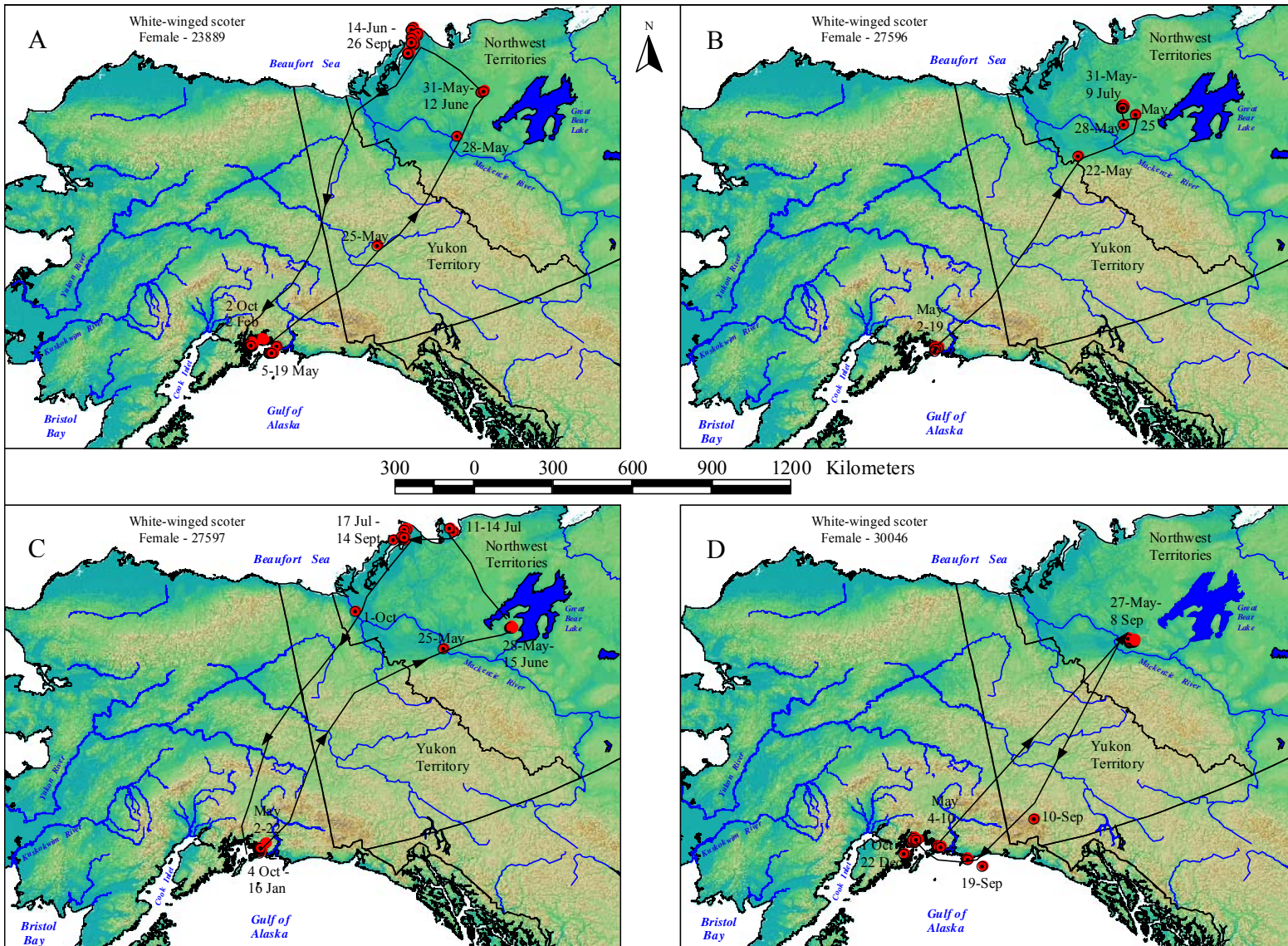


Figure 4. Movements of female white-winged scoters 23889, 27596, 27597, and 30046 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes.

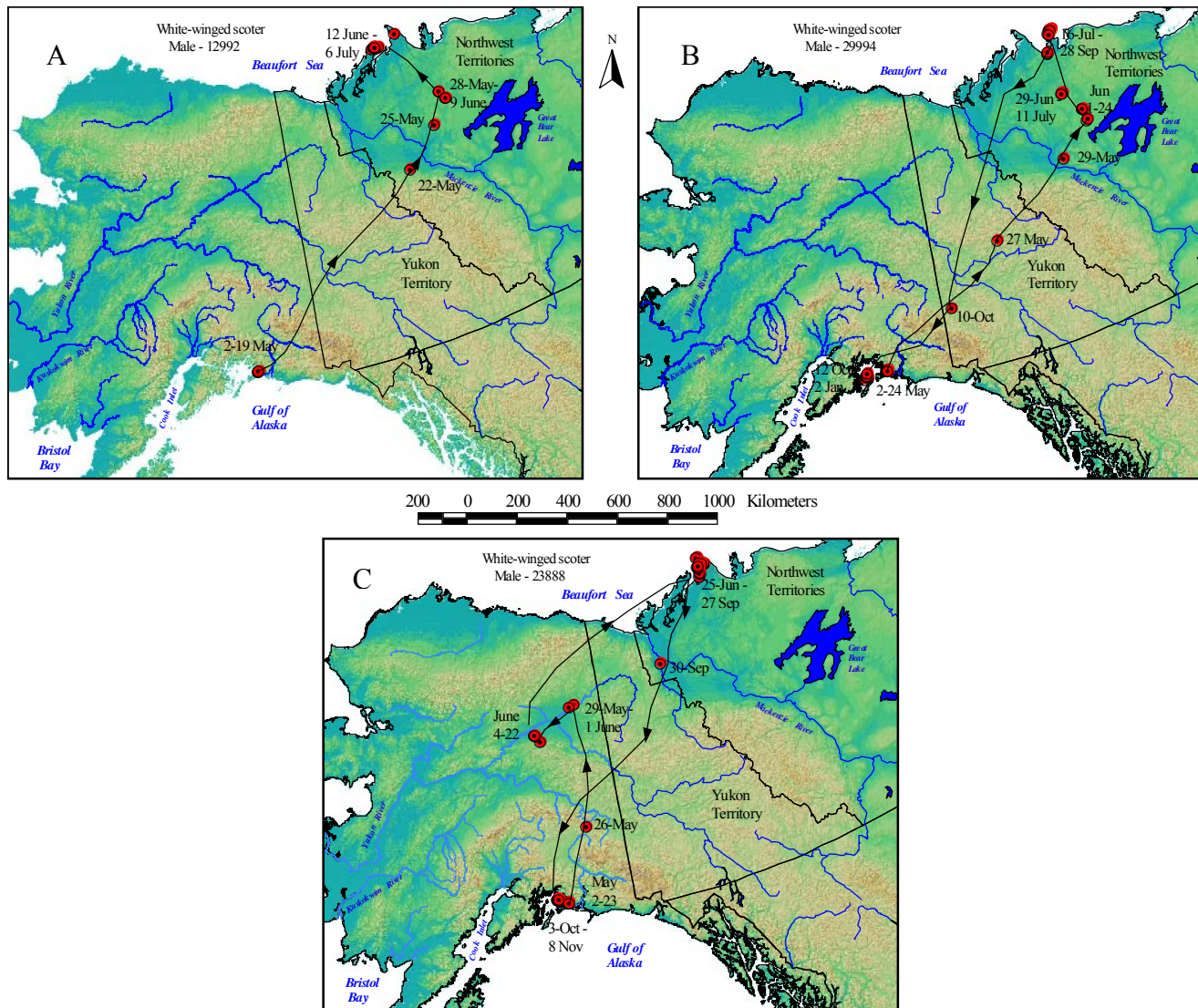


Figure 5. Movements of male white-winged scoters 12992, 29994, and 23888 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes.

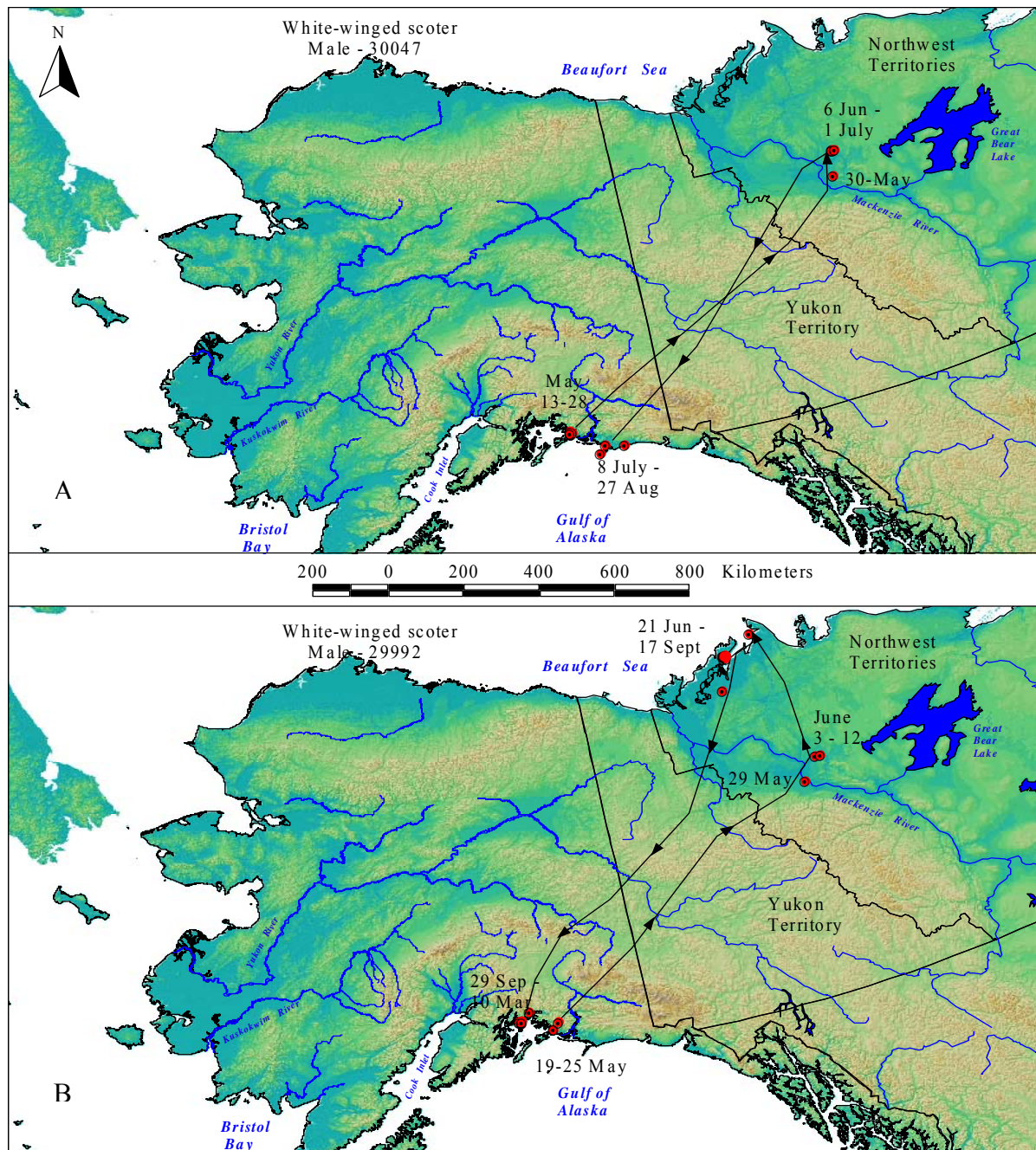


Figure 6. Movements of male white-winged scoters 30047 and 29992 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes.

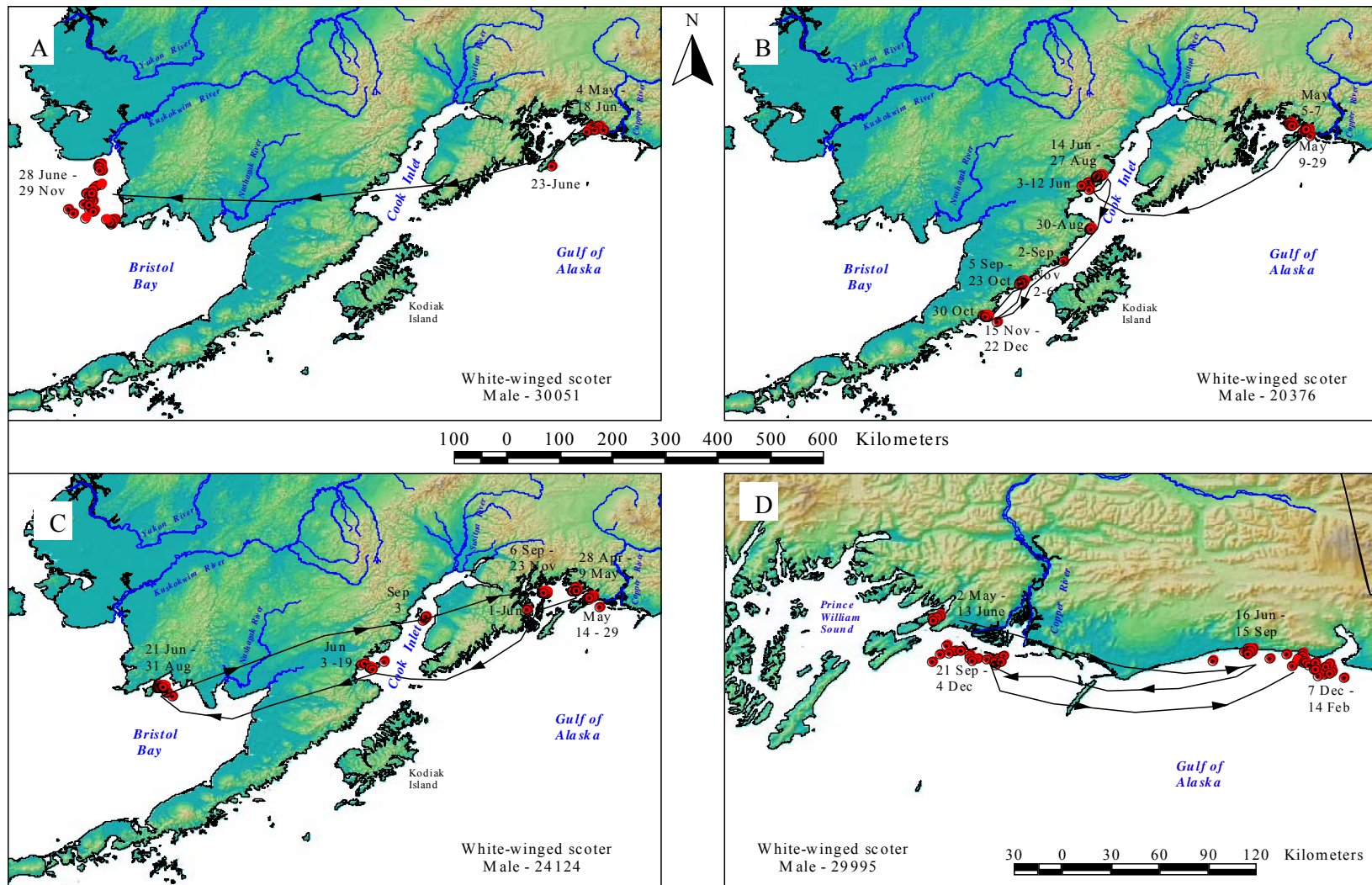


Figure 7. Movements of male white-winged scoters 30051, 20374, 24124 and 29995 marked with satellite transmitters from Prince William Sound, AK. Lines are not intended to depict actual routes

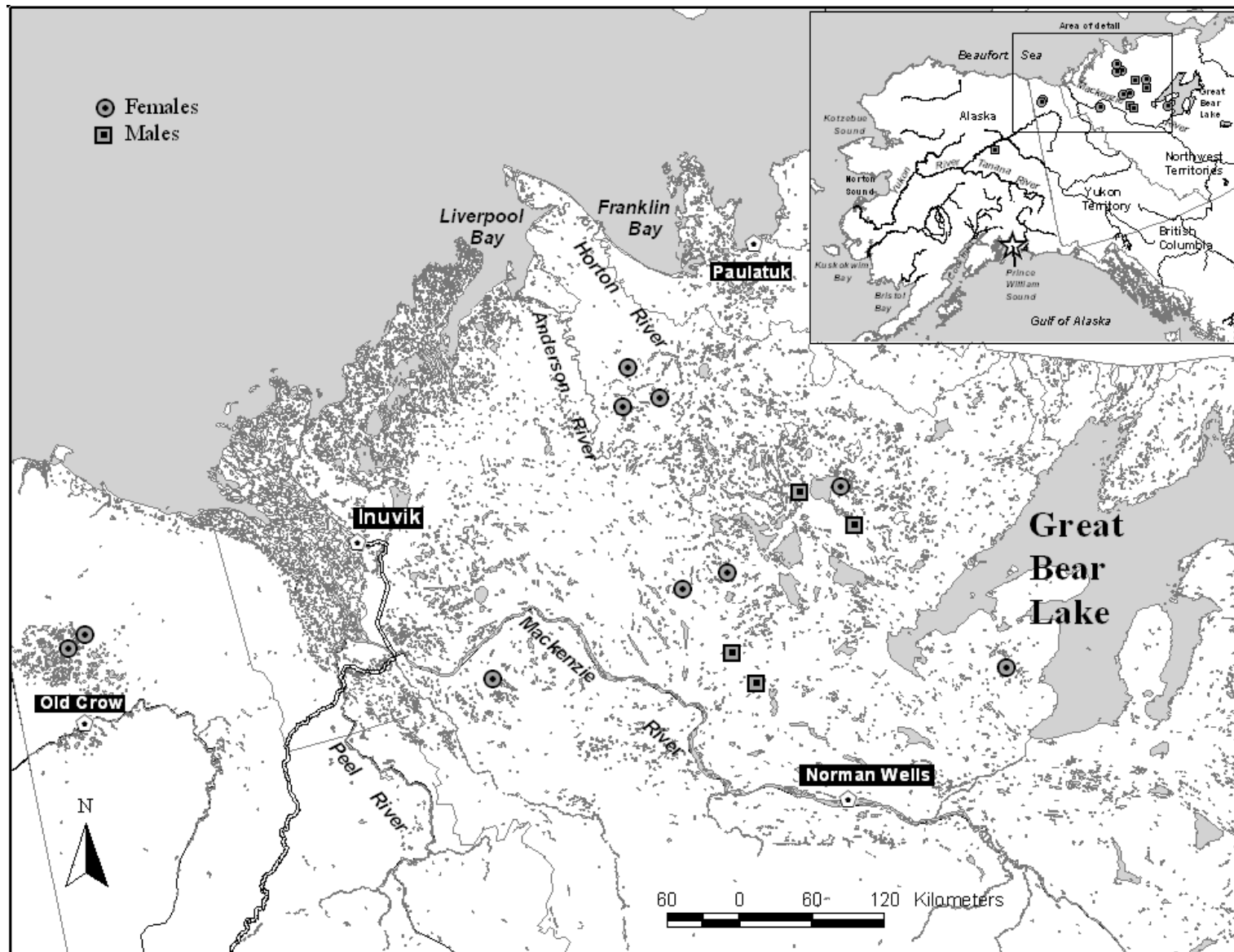


Figure 8. Map of breeding areas in Northwest and Yukon Territories, Canada used by white-winged scoters from Prince William Sound, Alaska. One male settled further east in the Yukon Flats, Alaska and is not depicted.

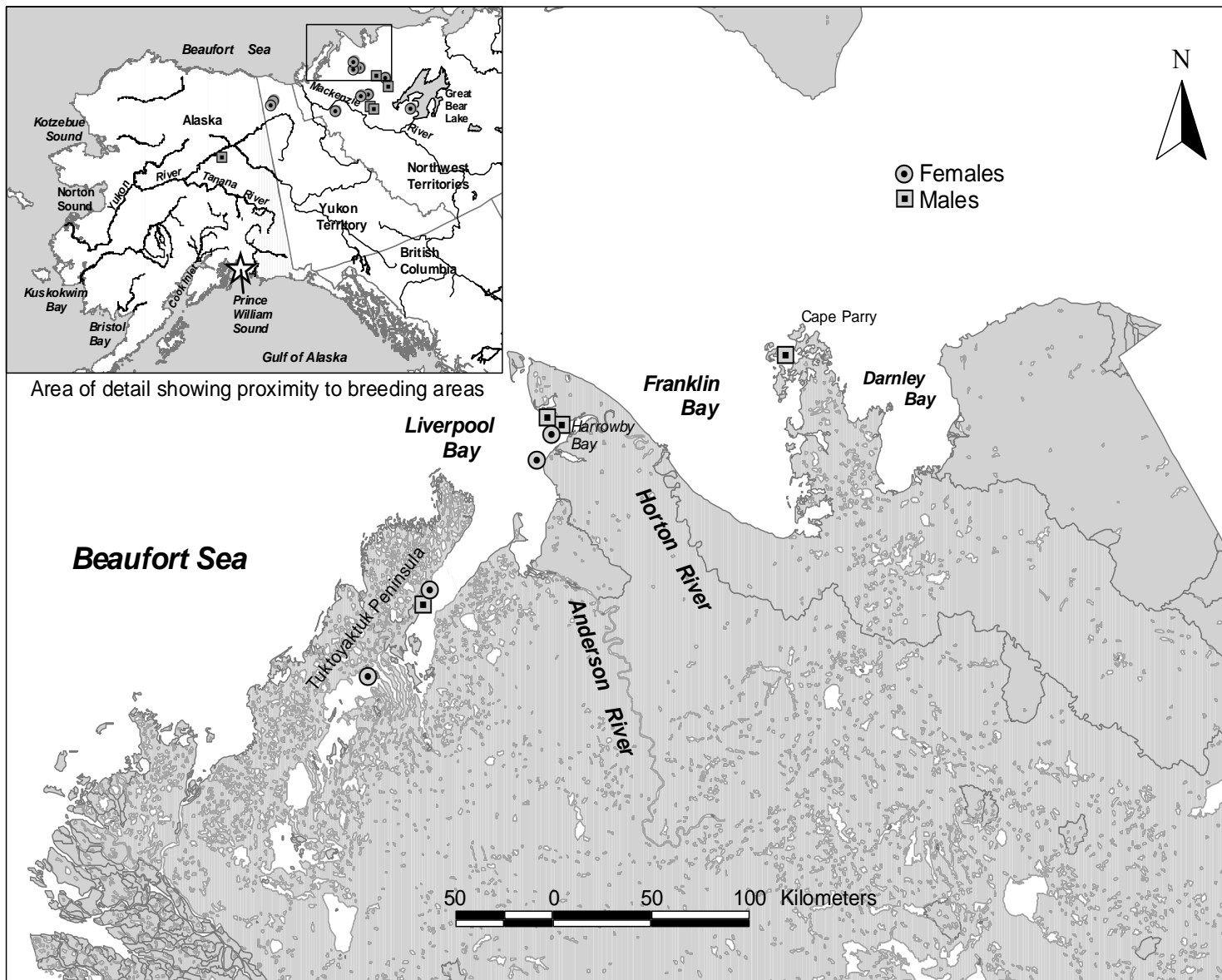


Figure 9. Map of molting areas in the Beaufort Sea, Northwest Territories, Canada used by white-winged scoters from Prince William Sound, Alaska. Additional molting sites were located in the Bering Sea and Gulf of Alaska.